

# Renewable Agriculture and Food Systems

#### cambridge.org/raf

# **Preliminary Report**

Cite this article: Candelaria-Morales NP, Grossman J, Fernandez A, Rogers M (2022). Exploring multifunctionality of summer cover crops for organic vegetable farms in the Upper Midwest. Renewable Agriculture and Food Systems 37, 198–205. https://doi.org/10.1017/ S1742170521000545

Received: 31 March 2021 Revised: 4 November 2021 Accepted: 29 November 2021

First published online: 14 February 2022

#### Keywords

Beneficial arthropods; ecosystem services; Midwest; nutrients; summer cover crops

Author for correspondence: Naomy P. Candelaria-Morales, E-mail: cande036@umn.edu

© The Author(s), 2022. Published by Cambridge University Press. This is an Open Access article, distributed under the terms of the Creative Commons Attribution licence (http://creativecommons.org/licenses/by/4.0/), which permits unrestricted re-use, distribution and reproduction, provided the original article is properly cited.



# Exploring multifunctionality of summer cover crops for organic vegetable farms in the Upper Midwest

Naomy P. Candelaria-Morales (a), Julie Grossman (b), Adria Fernandez (b) and Mary Rogers (b)

Department of Horticulture, University of Minnesota, St. Paul, MN, USA

#### Abstract

Intensive vegetable crop rotations can have detrimental effects on soil health, draining soil of organic matter reserves and necessitating nitrogen (N) inputs. In addition, many vegetable crop rotations leave little time or space to integrate beneficial arthropod and pollinator habitat into crop rotations; the lack of habitat may cause declines in beneficial arthropods, which can lead to insufficient pollination services and increased pest pressure. Nine treatments, each containing one to seven species of cover crops, were evaluated for flowering, aboveground biomass production and N content, soil NO3-N contribution after biomass incorporation, and beneficial arthropod visitation. A seven-species mix composed of oat (Avena sativa L.), field pea (Pisum sativum subsp. Arvense L.) and five clover species (Trifolium spp.) added the largest amount of biomass (8747 kg ha<sup>-1</sup>). Likewise, this mix contributed the most organic N (265.6 kg N  $\mathrm{ha}^{-1}$ ), and increased soil NO $_{3}^{-}$ -N after biomass incorporation (10.9 mg NO<sub>3</sub>-N kg<sup>-1</sup> of soil). Buckwheat (Fagopyrum esculentum Moench) and phacelia (Phacelia tanacetifolia Bendth.) monoculture produced most abundant floral resources. Beneficial arthropods observed included pollinators (native, honey and bumblebees), predators (syrphid flies and green lacewings) and parasitoids. Increased floral diversity was associated with abundance of flies in the Syrphidae family. Phacelia monoculture was most attractive for bees in the Apidae and Halictidae family, both of which may provide pollination services. These results highlight floral visitation patterns as an indicator for beneficial insect community support and conservation, especially in summer months, when greater insect reproduction occurs. Summerplanted cover crops are an underexplored rotation option for organic farming systems in the Upper Midwest, and may provide a wide range of ecosystem services including increases in available soil N and beneficial arthropod populations.

#### Introduction

Intensification of agricultural production systems can lead to natural resource depletion and reduced biodiversity. Although intensively managed agroecosystems can increase food production in the short-term, longer-term outcomes include loss of natural habitats and biodiversity of flora and fauna, water contamination, soil degradation and reduced capacity for crop productivity (Tilman *et al.*, 2002; Zhang *et al.*, 2015; Imadi *et al.*, 2016). In particular, low-diversity rotations can lead to soil fertility imbalances and declines in beneficial arthropod habitat, indirectly affecting soil degradation and water pollution (Büchs *et al.*, 1997; Pender and Mertz, 2006; O'Rourke *et al.*, 2008; Bowles *et al.*, 2020). Vegetable systems require high nutrient inputs for sustained yields, which may result in overfertilization and surface water eutrophication (Agostini *et al.*, 2010). Simplification of these systems limits the availability and variety of floral resources in Midwestern farming landscapes (Wäckers *et al.*, 2007; Blaauw and Isaacs, 2014; Forister *et al.*, 2019). This, in turn, reduces the ability of the land-scape to support beneficial arthropods (O'Connor, 2013).

Organic farming practices can reduce soil erosion and increase soil organic matter (SOM), thereby enhancing soil structure and stability (Pulleman *et al.*, 2000; Leithold *et al.*, 2014; Seitz *et al.*, 2019). These outcomes are often achieved via the use of cover crops to meet certified organic system diversification requirements (NOP, 2021), presenting an opportunity for inclusion of a range of cover crop species that meet the dual goals of soil and beneficial insect habitat improvement. This is especially true in Upper Midwest vegetable systems, where a variety of short fallow periods exist between cropping cycles when cover crops could be integrated (Wauters *et al.*, 2021). For example, cover crops could be established in spring, maturing during summer, and terminated prior to fall crop planting, or after an early spring cash crop, maturing throughout the late summer months. Upper Midwest summers are increasingly hot and humid with frequent rain events, which increase summer soil erosion and nutrient leaching (NOAA, 2013; Sumiahadi *et al.*, 2019). For organic vegetable farmers in northern regions,

cover crops planted during the summer months have the potential to reduce negative impacts of these climate shifts while increasing a range of ecosystem services, including production of aboveground biomass, biomass N, suppression of weeds and floral resources for pollinators (Creamer and Baldwin, 2000; Wilson *et al.*, 2018).

Cover crop species planted in the summer months, even in the coldest US climates, could provide the multifunctionality needed to meet numerous agroecosystem goals. Cover crops supply critical nutrients to organic vegetable production (Parr et al., 2011; Schipanski et al., 2014; Blesh, 2017; Blesh et al., 2019; Piotrowska-Długosz and Wilczewski, 2020). They have been shown to contribute up to 300 kg N ha<sup>-1</sup> (Creamer and Baldwin, 2000; Fageria et al., 2005; Sarrantonio, 2007; Parr et al., 2011; Asik et al., 2020), and increase soil fertility through changes in physical, chemical and biological parameters (Drinkwater et al., 1998; Reicosky and Forcella, 1998; Daryanto et al., 2018). Moreover, when cover crop functional groups are combined, greater multifunctionality can be obtained (Finney et al., 2017; Kaye et al., 2019). For example, when legumes and grasses are combined, legumes may fix more N due to competition with grasses in acquiring soil N (Jensen, 1996; Høgh-Jensen & Schjoerring, 2001; Blesh, 2017). These crops also offer an opportunity to add floral resources throughout the growing season. Adding flowering cover crops to the rotation supports diversity and abundance of beneficial arthropods, including bees and natural enemies of pest (O'Connor, 2013; Redlich et al., 2018). These arthropods can provide ecosystem services to the farm, including pollination and predation and parasitism of pest (Hooks et al., 1998; Landis et al., 2005).

Data are scarce regarding the degree to which summer-planted flowering cover crops can support beneficial arthropod communities while simultaneously supporting soil health services, particularly soil nutrient cycling, in vegetable production systems. The objectives of this research were to assess possible contributions of 16 cover crop species and species mixes toward enhancement of (1) soil health factors (biomass production and C:N ratio content,  $NO_3^-$ -N contributions) and (2) beneficial arthropod habitat (population abundance, diversity and richness).

# Materials and methods

#### Experimental location and treatments

A summer-planted cover crop trial was established in a single growing season from May to September 2019 at the University of Minnesota Agricultural Experimental Station (MAES), Twin Cities campus, St. Paul, MN (44.990829, -93.175832). Species were selected to represent commonly used Upper Midwest cover crops and their expected contribution of beneficial arthropod floral resources and nutrient cycling. Results from an on-farm preliminary trial and recommendations from the Xerces Society for Invertebrate Conservation and by farmer collaborators further supported species selection (Table 1). Regional mean annual rainfall is 813 mm, and the soil is classified as a Waukegan silt loam (fine-silty over sandy or sandy-skeletal, mixed, superactive, mesic Typic Hapludolls). Treatments were arranged in a randomized complete-block experimental design with four replications, with each treatment plot measuring 3 m<sup>2</sup>. Cover crops were broadcast sown at two dates (early and late), lightly raked to increase seed to soil contact, and terminated at the time of seed formation (Table 2). Previous fertility management prior to experiment establishment included historical manure additions, as well as regular

Table 1. Treatment species and seeding rates

Treatment	Common (scientific) name	Seeding rate (kg ha <sup>-1</sup> )				
В	Buckwheat ( <i>Fagopyrum esculentum</i> Moench)	113				
Ph	Phacelia ( <i>Phacelia tanacetifolia</i> Bendth.)	14				
S	Sunflower (Helianthus annuus L.)	11				
Sh	Sunn hemp (Crotalaria juncea L.)	23				
BAC	Buckwheat, annual ryegrass (Lolium multiflorum L.), and clover spp. [crimson clover (Trifolium incarnatum L.), berseem clover (T. alexandrinum L.), medium red clover (T. pratense L.), ladino white clover (T. repens L.) and alsike clover (T. hybridum L.))	23, 23, 2ª				
BP	Buckwheat and partridge pea (Chamaecrista fasciculata Michx.)	79, 17				
BPhS	Buckwheat, phacelia and peredovik sunflower	19, 3, 2				
MD	Buckwheat, canola ( <i>Brassica napus</i> L.), oilseed radish ( <i>Raphanus sativus</i> L.), field peas, pearl millet ( <i>Pennisetum glaucum</i> L.) and oats	8, 1, 1, 17, 3, 11				
OFC	Oats (Avena sativa L.), field peas (Pisum sativum subsp. Arvense L.) and clover spp.	113, 113, 2 <sup>a</sup>				

<sup>&</sup>lt;sup>a</sup>Consistent for all clover species.

Table 2. Dates of field operations

	Date						
Operation	2018						
Early cover crop planting	1 May						
Late cover crop planting	15 May						
Flower counts	26 Jun/3, 9, 16, 23, 30 Jul/6 Aug						
Insect counts	26 Jun/9, 11, 16, 23, 30 Jul/3, 6 Aug						
Sampled biomass	15, 29 Jul/13 Aug						
Mowed cover crops	15, 29 Jul/13 Aug						
Sampled soil (post-mowed)	15, 29 Jul/13 Aug						
Tilled biomass	31 Jul/19, 29 Aug						
Sampled soil post-tillage	14 Aug/3, 10 Sep						

additions of composted turkey litter. No synthetic or organic fertilizers and pesticides were applied in the 2 years prior to the experiment, or during the experimental growing season.

#### Biomass and soil sampling

Plant biomass was collected immediately before cover crop termination (Table 2) from a 0.5 m<sup>2</sup> sampling area, 0.3 m away from plot borders. To maximize flowering time yet avoid seed-set and cover crops reseeding, cover crops with early [buckwheat (*Fagopyrum esculentum* Moench)] and medium-length [phacelia monoculture (*Phacelia tanacetifolia* Bendth), and oat (*Avena* 

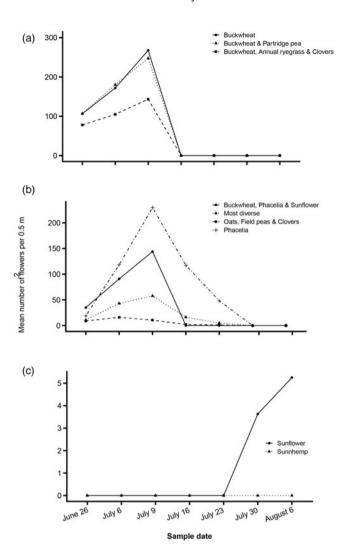
sativa L.), field pea (Pisum sativum subsp. Arvense L.) and clover (Trifolium incarnatum L., T. alexandrinum L., T. pratense L., T. repens L., T. hybridum L.) mix] flowering were terminated on July 15, sunn hemp (Crotalaria juncea L.) was terminated on July 29 due to low biomass production and no flowering, and lateflowering crops [sunflower (Helianthus annuus L.)] were terminated on August 13. Aboveground biomass samples were categorized and sorted as cover crops and weeds, then oven-dried at 60°C for at least 48 h. Samples were then weighed, and ground to 2 mm for C:N ratio determination. One sunn hemp treatment plot had no biomass collected due to low germination; this plot was excluded from analysis. Cover crop termination was performed with a walkbehind tractor (BCS) and flail mower attachment. Due to unfavorable weather conditions and wet soils, biomass could not be incorporated in the soil with rotary tillage until 2 weeks after termination. Soils were sampled at cover crop termination and again approximately 2 weeks following tillage (Table 2). Ten soil cores were collected to 15 cm depth, using 2.5 cm diameter stainless steel soil probes, and homogenized. Samples were stored in paper bags and oven-dried at 35°C for at least 48 h. Dry soil samples were ground to 2 mm for soil inorganic N (NO<sub>3</sub>-N) analysis. Inorganic NO<sub>3</sub>-N was measured using a KCl extraction (modified from Robertson et al., 1999). Briefly, 8 g of dry soil was weighed into 50 mL centrifuge tubes, 40 mL 1 M KCl was added, and tubes were shaken for 1 h at approximately 240 rpm. Samples were allowed to settle for at least 1 h, and supernatant was filtered through #1 Whatman papers and collected in scintillation vials. Vials were frozen for analysis, which was later performed using a colorimetric nitrate assay (Doane and Horwath, 2003). Biomass was analyzed for C and N content using a dry combustion analyzer (Elementar VarioMax CN analyzer).

## Floral and arthropod sampling

Flower units, measurement used to quantify flower abundance, for each individual species were quantified and recorded within a 0.5 m<sup>2</sup> area, beginning when ten or more open inflorescence were observed within each plot (Table 2). Flower units were categorized depending on the species. For buckwheat and phacelia, each bifurcated stem was considered one flower unit. Flower units for sunflower, clovers and field peas were determined by each inflorescence. Cover crop peak blooms were categorized as early, early-mid and late bloom time (Fig. 1). Plots were terminated when buckwheat started producing mature seeds (i.e., darkcolored filled pods), even if other species from the treatment were not in bloom (Table 2). Beneficial arthropod communities were sampled every 7-8 days by sampling the perimeter of each plot with a sweep net for 60 s. Sampling did not occur the week of July 7–17 due to weather conditions. After collection, arthropods were preserved in a -20°C freezer until family-level taxonomic identification, and preserved in an 80% ethanol solution thereafter. Specimens were identified to family using the keys provided by Triplehorn et al. (2005) and Marshall (2006) as well as photographs and descriptions available in www.BugGuide.net. Selected specimen identifications were verified by Dr Ralph Holzenthal, Department of Entomology, University of Minnesota. Beneficial arthropod abundance, diversity and richness were assessed and categorized by functional groups for all flowering cover crops.

#### Statistical analysis

For soil and cover crop biomass, we report dry biomass quantity  $(kg ha^{-1})$ , biomass N content  $(kg N ha^{-1})$ , and soil N  $(NO_3^--N)$ 



**Fig. 1.** Mean numbers of open flowers across time on treatment plants that reached peak bloom (a) early, (b) mid-early and (c) late.

at 0 and 2 weeks after biomass incorporation. Changes reported in NO<sub>3</sub>-N were calculated by subtracting mean pre-termination soil NO<sub>3</sub>-N values from those taken 2 weeks following termination. Because soil conditions (temperature and moisture) are known to impact N mineralization rates, and times of sampling and termination in this study varied by flowering times (Table 2), reported soil NO<sub>3</sub>-N values should be considered as a potential of summer cover crops, and not absolute. Cover crop treatment fixed effects on soil properties and biomass production (P < 0.05) were analyzed using a mixed-effects model ANOVA with replication as a random effect (package lme4; Bates et al., 2015). Floral, beneficial arthropod and predominant group taxa abundance as a function of cover crop treatments were analyzed using a generalized linear model with a Poisson family distribution (package R; R Core Team, 2019). Beneficial arthropod data are reported as the total number and relative abundances of collected families by treatment. To assess the diversity of pollinators and natural enemies, for each experimental unit, Shannon indices were calculated using insect family as a proxy for species (R vegan package; Okansen et al., 2019). Linear mixed-effects models with terms for cover crop treatment as the main effect and replicated plot as a random effect were used to compare richness and diversity of natural enemies among

Table 3. Mean C:N ratio, cover crop and weed aboveground biomass, aboveground biomass N and percent N by treatment

			Cov		Weed				
Cover crop treatment	C:N	Biomass kg N ha <sup>-1</sup>	Bi	omass N kg N ha <sup>-1</sup>	Soil inorg Before	After	C:N	Biomass kg N ha <sup>-1</sup>	Biomass N
В	31.1 bc	2730 cd	1.5	43.1 bc	1.3 b	8.1 b	25.1 b	2295 b	39.0 bc
BAC	24.8 c	3847 bc	1.8	70.4 b	1.3 b	7.5 bc	29.8 b	1330 bc	19.5 с
BP	37.3 ab	1955 d	1.2	23.7 bc	1.3 b	10.6 ab	25.7 b	2025 b	34.1 bc
BPhS	31.8 b	2386 cd	1.3	31.1 c	1.3 b	11.1 a	26 b	2115 bc	35.6 bc
MD	17.3 d	1866 d	2.8	53.5 bc	1.6 ab	8.2 b	25.9 b	2600 b	45.3 b
OFC	16.0 d	8747 a	2.9	265.6 a	2.5 a	13.5 a	18.3 c	377 c	8.7 c
Ph	34.3 b	5035 b	1.2	61.7 bc	1.2 b	7.5 bc	23.7 bc	691.5 c	12.1 c
S	42.3 a	4250 bc	1.1	46.9 bc	0.5 c	5.5 c	41.2 a	3760 ab	44.4 bc
Sh	20.9 cd	20 e	2.2	0.6 d	1.2 b	6.1 bc	30.2 b	5735 a	84.3 a

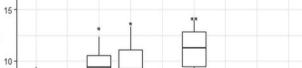
B, buckwheat; BAC, buckwheat, annual ryegrass and clovers; BP, buckwheat and partridge pea; BPhS, buckwheat, phacelia and sunflower; MD, buckwheat, canola, radish, field peas, pearl millet and oats; OFC, oats, field peas and clovers; Ph, phacelia; S, sunflower. Results followed by the same letter within each column are not different (P < 0.05).

treatments (package lme4; Bates et al., 2015). Response variables of biomass, biomass N, richness and Shannon's diversity indices were square root transformed to normalize error and satisfy the assumption of equal variance. All analyses were conducted with R, version 3.5.3 (The R Foundation for Statistical Computing, 2019) and R Studio, version 1.2.1 (R Studio Team, 2018).

### **Results and discussion**

Biomass production, C:N ratio and soil inorganic N

Cover crop biomass production (kg ha<sup>-1</sup>) varied across treatments (P < 0.001) with higher-biomass treatments trending toward higher N-contributing treatments. The oat, field pea and clover (OFC) treatment produced the greatest biomass (8747  $\pm$  1600 kg ha<sup>-1</sup>) and suppressed the most weeds (Table 3). This treatment also produced the highest biomass N ( $265 \pm 73 \text{ N kg ha}^{-1}$ ), attributed to its high biomass and low C:N ratio. Clover biomass did not reflect the original proportion of clover in the OFC mix, resulting in an over-representation of oats and field peas in this treatment. Although many clovers are slower to establish than either field pea or oat, and often do not flower the first year, crimson clover flowered by week 3. As expected, weed biomass was negatively correlated to cover crop biomass (P < 0.001,  $R^2 = -0.63$ ). This result supports previous findings that summerplanted cover crops can produce sufficient biomass to be competitive with weed species (Finney et al., 2016; Sharma et al., 2018; Stepanovic et al., 2018). Sunn hemp, a buckwheat and partridge pea mix (BP), and the most diverse plant family mix (MD) had low cover crop biomass and correspondingly greater amounts of weeds due to poor germination relative to other treatments. Sunn hemp is a warm-climate cover crop that is extensively used in the southern US and the Caribbean (Halbrendt, 2010; Morris et al., 2015). Sunn hemp germination may have been affected by cooler air and soil temperatures during the planting dates of May 1 and May 15, when temperatures in Minnesota varied between 7 and 25°C. The mixture treatments, BP and MD, both included buckwheat, known to have rapid establishment and competitive germination, potentially affecting the



Increased NO<sub>3</sub> in soil after biomass incorporation

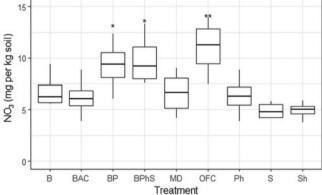


Fig. 2. Effect of cover crop treatments on inorganic N from 0 to 2 weeks after biomass incorporation. B, buckwheat; BAC, buckwheat, annual ryegrass and clovers; BP, buckwheat and partridge pea; BPhS, buckwheat, phacelia and sunflower; most diverse (MD); buckwheat, canola, radish, field peas, pearl millet and oats; OFC, oats, field peas and clovers; Ph, phacelia; S, sunflower; Sh, sunn hemp. Asterisk over the boxes denote statistically different means at \* $P \le 0.05$  and \*\* $P \le 0.001$ .

establishment of other species in the mix. However, the mix containing buckwheat, annual ryegrass and clovers (BAC) had comparable biomass to phacelia and sunflower, second and third highest biomass production species. Clover species' germination in BAC was not representative of seeding rate.

Soil NO $_3$ -N varied across treatments (P < 0.001), increasing from 26 to 55% 2 weeks following biomass incorporation in BP, BPhS and OFC treatments (Fig. 2). Although this is expected for OFC, which produced significantly higher biomass N that likely drove soil N increases, this result was surprising for BPhS and BP, as these treatments did not produce substantially greater amounts of biomass N relative to any other treatment. In these cases, available soil N could possibly have increased via N inputs not quantified in this study, including N from root residue decomposition (Bukovsky-Reyes et al., 2019).

202

Table 4. Season-long total number of individuals and relative abundance (RA) for each taxon encountered across nine treatments

Таха		В		BAC		ВР		BPhS		MD		OFC		Ph		S	
	Total	RA (%)	Total														
Coleoptera																	
Coccinellidae	-	-	3	3.2	-	-	2	3.4	32	21.8	-	-	-	-	-	-	37
Diptera																	
Syrphidae	67	76.1	74	77.9	46	67.6	41	69.5	97	66.0	31	58.5	96	72.7	2	7.1	454
Hemiptera																	
Anthocoridae	2	2.3	1	1.1	1	1.5	_	-	-	-	1	1.9	1	0.8	2	7.1	8
Reduviidae	2	2.3	-	-	-	-	-	-	-	-	-	-	-	-	-	-	2
Hymenoptera																	
Andrenidae	-	-	-	-	-	-	-	-	-	-	2	3.8	-	-	1	3.6	3
Anthophorini	-	-	-	-	-	-	-	-	-	-	-	-	-	-	5	17.9	5
Apidae	3	3.4	1	1.1	11	16.2	7	11.9	2	1.4	1	1.9	20	15.2	4	14.3	49
Braconidae	2	2.3	1	1.1	2	2.9	-	-	3	2.0	7	13.2	1	0.8	2	7.1	18
Colletidae	-	-	-	-	-	-	1	1.7	1	0.7	-	-	-	-	-	-	2
Encyrtidae	_	-	1	1.1	-	-	-	-	1	0.7	-	-	-	-	1	3.6	3
Eulophidae	5	5.7	5	5.3	4	5.9	2	3.4	7	4.8	4	7.5	1	0.8	2	7.1	30
Eurytomidae	-	-	1	1.1	-	-	1	1.7	1	0.7	-	-	-	-	-	-	3
Halictidae	1	1.1	2	2.1	-	-	2	3.4	-	-	-	-	12	9.1	1	3.6	18
Ichneumonidae	-	-	-	-	-	-	_	-	1	0.7	2	3.8	-	-	-	-	3
Pteromalidae	3	3.4	-	-	2	2.9	2	3.4	2	1.4	4	7.5	-	-	8	28.6	21
Sphecidae	-	-	1	1.1	-	-	-	-	-	-	-	-	-	-	-	-	1
Vespidae	-	-	-	-	-	-	-	_	-	_	1	1.9	-	-	-	_	1
Neuroptera																	
Chrysopidae	3	3.4	5	5.3	1	1.5	1	1.7	-	-	-	-	1	0.8	-	-	11
Odonata																	
Coenagrionidae	_	-	_	-	1	1.5	_	-	_	_	_	-	_	_	_	-	1
Total	88		95		68		59		147		53		132		28		670

B, buckwheat; BAC, buckwheat, annual ryegrass and clovers; BP, buckwheat and partridge pea; BPhS, buckwheat, phacelia and sunflower; MD, buckwheat, canola, radish, field peas, pearl millet and oats; OFC, oats, field peas and clovers; Ph, phacelia; S, sunflower.

Cover crop mixtures that include legumes or low C:N biomass species often increase soil health indicators, including SOM and N (Finney et al., 2016; Kaye et al., 2019). We observed 73-100% greater biomass N provided by the OFC treatment relative to all other treatments. However, other treatments that also included legumes with non-legume components (Table 1) did not produce comparable amounts of biomass N. These other treatments included buckwheat, known to have earlier and faster growth rates (Hogg et al., 2011), relative to other species included in the evaluated mixtures. This rapid growth of buckwheat may have decreased legume productivity and subsequent N contributions via competitive effects. Legumes generally contribute higher biomass N than buckwheat, with recorded levels of 217 kg ha<sup>-1</sup> compared to 21.3 kg ha<sup>-1</sup>, respectively (Marcinkonis et al., 2007; Parr et al., 2011), yet individual species' productivity in a given environment ultimately controls overall N contributions. Among the buckwheat-containing treatments, the BAC mix trended towards higher biomass N, but was not statistically different from B, BP, MD treatments (P > 0.05). The MD treatment had 51-79% lower biomass production compared to OFC and BAC mixes, and to phacelia and sunflower monocultures, also likely attributed to competition with mixture components (Sarrantonio, 2007). Results support that selecting appropriate competitive plant traits for mixes, especially favoring legumes, may present a challenge, as greater competitive establishment may benefit weed control but decrease germination of other mix species.

#### Floral resources

Buckwheat treatments started flowering extensively 42 days after planting date. Buckwheat monoculture produced the highest mean floral peak (flower units =  $374 \pm 95.9$ ) relative to other treatments containing buckwheat (ranging from  $56.2 \pm 14.1$  to  $202 \pm$ 71.6). Phacelia began flowering progressively 42 days after planting date, reaching peak bloom by July 9 (181  $\pm$  53.4), and culminating July 30. Although both buckwheat and phacelia monocultures had started flowering by the first collection sample, buckwheat in mixed culture flowered 14 days earlier than phacelia, providing early floral resources for a wide range of arthropod visitors. The OFC treatment started flowering 42 days after planting date and reached an average peak bloom of 11.2 ± 2.9 flower units. The sunflower treatment had a period of 76 days to flower from planting date, lasting 14 days and with an average peak bloom of  $8.9 \pm 1.6$  flower units. The sunn hemp treatment did not provide flowers during the experimental period.

#### Beneficial arthropod abundance, diversity and richness

Across all treatments, a total of 2668 arthropod specimens were collected from cover crop treatments, of which 25% (667) were categorized as beneficial arthropods, including natural enemies (predators and parasitoids) and pollinators (bees and hoverflies). The 667 beneficial specimens were subset into 18 taxonomic families with Syrphidae (68.1%) and Apidae (7.3%) comprising approximately 75% of all specimens collected, followed by Coccinellidae (5.5%) and Eulophidae (4.5%). Syrphidae had the highest relative abundance (77.9%), followed by Coccinellidae (21.8%) and Apidae (16.2%) (Table 4).

Abundance of beneficial arthropods in the OFC and sunflower treatments were lower than all other treatments, likely attributed to low inflorescence production in both treatments. The MD

treatment had the highest abundance of total beneficial arthropods (147) relative to both OFC (53) and S (28) treatments (Table 4). However, this was not statistically different from buckwheat, BAC, BP, BPhS and phacelia treatments.

Plant functional diversity has been correlated with higher arthropod diversity in a perennial savannah grassland (Haddad et al., 2001). However, our results demonstrate that higher plant functional diversity does not necessarily result in higher arthropod diversity. Although the MD treatment had the greatest beneficial arthropod abundance, it had lower richness than the monoculture or bi-culture treatments. Buckwheat (diversity/richness, respectively;  $0.66 \pm 0.14/2.6 \pm 0.26$ ) and BP ( $0.67 \pm 0.17/2.6 \pm 0.42$ ) treatments had a higher richness than the MD ( $0.22 \pm 0.08/1.3 \pm 0.28$ ) and phacelia ( $0.26 \pm 0.10/1.5 \pm 0.29$ ) treatment. B, BAC and BP displayed higher richness but no significant difference in diversity relative to OFC ( $0.35 \pm 0.11/1.4 \pm 0.29$ ) treatment. BAC ( $0.68 \pm 0.20/3.0 \pm 0.60$ ) treatment showed difference in diversity compared to sunflower ( $0.37 \pm 0.12/1.9 \pm 0.19$ ), however no difference in richness.

Members of the Syrphidae family are known to be attracted to small, white flowers such as buckwheat, Lobularia (alyssum) spp. and Coriandrum sativum (Hogg et al., 2011; Martinez-Uña et al., 2014). Considering these specimens composed 68% of the total amount of beneficial arthropods collected, the higher amounts of beneficial arthropods in treatments containing buckwheat (Table 4) are consistent to previous studies and throughout our treatments. Additionally, *Phacelia* sp. was particularly attractive to native and non-native pollinator groups such as Bombus, Halictidae and honeybees (Apis mellifera) (Table 4). The sunflower treatment flowered the latest and had short flowering length, explaining the low abundance of arthropods captured. However, arthropod diversity in sunflower was equivalent to longer-flowering treatments, demonstrating the capacity of sunflowers to attract a wide range of beneficial arthropods even during a shorter flowering period. Native species are believed to have co-evolved with endemic natural enemies (James et al., 2014). Therefore, sunflowers, which are native to North America, may play a key role in attracting and sustaining endemic beneficial arthropod populations (Fiedler and Landis, 2007).

#### **Conclusion**

This research explored the effect of summer-planted flowering cover crop mixes on concurrent N provision and presence of floral resources for beneficial arthropods. Nine treatments containing one to seven species of cover crops were evaluated for flowering time, N content, soil N contribution and beneficial arthropod attraction. The oat, field pea and clover mix added the most biomass, followed by phacelia and sunflower monocultures. The OFC treatment also provided the greatest increase in soil NO<sub>3</sub>-N from 0 to 2 weeks after biomass incorporation relative to all other treatments. Buckwheat monoculture and mixes demonstrated the shortest time to flower and ranked higher in all three arthropod indices: abundance, diversity and richness. In the Upper Midwest, a short summer fallow period in organic vegetable crop rotations presents a valuable opportunity to introduce cover crop mixes and provide additional ecosystem services. Such mixes typically include a variety of grasses, legumes and brassicas. Expanding cover crop options to include nontraditional flowering species attractive to beneficial insects, such a phacelia, may increase the provision of desired ecosystem services. As these flowering cover crop mixtures are introduced

into farmer's rotations, pollinator presence may also be increased. More research is needed to determine optimal termination timing, given that longer growth is needed for flowering, but the physiological changes associated with flowering could lead to less biomass-N availability (Schiltz *et al.*, 2005). This trade-off between ecosystem services is especially challenging to manage when using cover crop mixes. A mix with varying bloom times can extend the degree of pollinator habitat provision, yet leads to inconsistent biomass quality among mix species, as well as potential termination time challenges, as some species may go to seed prior to other species. Further research of ecosystem services trade-offs such as this is required to obtain greater clarity of their interactions and to what extent summer-planted cover crops may supply multifunctionality.

**Acknowledgments.** We would like to thank the Xerces Society and Sogn Valley Farm for selecting the cover crop treatment species, all the members of our labs for final commentaries and our family and friends. This work was supported by NRCS Conservation Innovation Project (NR203A750008G002) and North Central Region SARE (NC-SARE Project LNC19-423).

Conflict of interest. None.

#### References

- Agostini F, Tei F, Silgram M, Farneselli M, Benincasa P and Aller MF (2010) Decreasing nitrate leaching in vegetable crops with better N management. Genetic engineering, biofertilisation. Soil Quality and Organic Farming 4, 147–200.
- Asik BB, Uzun A and Acikgöz E (2020) Seeding rate and cultivar impacts on nutrient uptake of field pea under fertile soil condition. Chilean Journal of Agricultural Research 80, 11–20.
- Bates D, Mächler M, Bolker B and Walker S (2015) Fitting linear mixed effects models using lme4. *Journal of Statistical Software* 67, 1–48.
- Blaauw BR and Isaacs R (2014) Flower plantings increase wild bee abundance and the pollination services provided to a pollination-dependent crop. *Journal of Applied Ecology* **51**, 890–898.
- Blesh J (2017) Functional traits in cover crop mixtures: biological nitrogen fixation and multifunctionality. *Journal of Applied Ecology* **2017**, 1–11.
- Blesh J, VanDusen BM and Brainard DC (2019) Managing ecosystem services with cover crop mixtures on organic farms. *Agronomy Journal* 111, 826–840.
- Bowles TM, Mooshammer M, Socolar Y, Calderón F, Cavigelli MA, Culman SW, Deen W, Drury CF, Garcia y Garcia A, Gaudin ACM, Harkcom WS, Lehman RL, Osborne SL, Robertson GP, Salerno J, Schmer MR, Strock J and Grandy AS (2020) Long-term evidence shows that crop-rotation diversification increases agricultural resilience to adverse growing conditions in North America. *One Earth* 2, 284–293.
- Büchs W, Harenberg A and Zimmermann J (1997) The invertebrate ecology of farmland as a mirror of the intensity of the impact of man?— An approach to interpreting results of field experiments carried out in different crop management intensities of a sugar beet and an oil seed rape rotation including set-aside. *Biological Agriculture and Horticulture* 15, 83–107.
- Bukovsky-Reyes S, Isaac ME and Blesh J (2019) Effects of intercropping and soil properties on root functional traits of cover crops. Agriculture, Ecosystems and Environment 285, 1–10.
- Creamer NG and Baldwin KR (2000) An evaluation of summer cover crops for use in vegetable production systems in North Carolina. *HortScience* 35, 600–603.
- Daryanto S, Fu B, Wang L, Jacinthe PA and Zhao W (2018) Quantitative synthesis on the ecosystem services of cover crops. *Earth-Science Reviews* 185, 357–373.
- **Doane TA and Horwath WR** (2003) Spectrophotometric determination of nitrate with a single reagent. *Analytical Letters* **36**, 2713–2722.

- **Drinkwater LE, Wagoner P and Sarrantonio M** (1998) Legume-based cropping systems have reduced carbon and nitrogen losses. *Nature* **396**, 262–265.
- Fageria NK, Baligar VC and Bailey BA (2005) Role of cover crops in improving soil and row crop productivity. *Communications in Soil Science and Plant Analysis* 36, 2733–2757.
- Fiedler AK and Landis DA (2007) Attractiveness of Michigan native plants to arthropod natural enemies and herbivores. *Environmental Entomology* 36, 751–765.
- **Finney DM, White CM and Kaye JP** (2016) Biomass production and carbon/nitrogen ratio influence ecosystem services from cover crop mixtures. *Agronomy Journal* **108**, 39–52.
- Finney DM, Murrell EG, White CM, Baraibar B, Barbercheck ME, Bradley BA, Cornelisse S, Hunter MC, Kaye JP, Mortensen DA, Mullen CA and Schipanski M (2017) Ecosystem services and disservices are bundled in simple and diverse cover cropping systems. Agriculture and Environmental Letters 2, 170033.
- Forister ML, Pelton EM and Black SH (2019) Declines in insect abundance and diversity: we know enough to act now. *Conservation Science and Practice* 1, 1–8.
- Haddad NM, Tilman D, Haarstad J, Ritchie M and Knops JMH (2001) Contrasting effects of plant richness and composition on insect communities: a field experiment. *American Naturalist* 158, 17–35.
- Halbrendt JC (2010) Here comes the sunn: a comprehensive analysis of the cover crop sunnhemp (*Crotalaria juncea* L.) and its potential for seed production and weed control in Puerto Rican agricultural systems. University of Puerto Rico, Mayaguez.
- Hogg BN, Bugg RL and Daane KM (2011) Attractiveness of common insectary and harvestable floral resources to beneficial insects. *Biological Control* 56, 76–84.
- Høgh-Jensen H and Schjoerring JK (2001) Rhizodeposition of nitrogen by red clover, white clover and ryegrass leys. Soil Biology and Biochemistry 33, 439–448.
- **Hooks CRR, Valenzuela HR and Defrank J** (1998) Incidence of pests and arthropod natural enemies in zucchini grown with living mulches. *Agriculture Ecosystem and Environment* **69**, 217–231.
- Imadi SR, Shazadi K, Gul A and Hakeem KR (2016) Sustainable Crop Production System. In Hakeem K, Akhtar M and Abdullah S (eds), In Plant, Soil and Microbes. Cham: Springer, pp. 103–116.
- James DG, Seymour L, Lauby G and Buckley K (2014) Beneficial insects attracted to native flowering buckwheats (*Eriogonum Michx*) in central Washington. *Environmental Entomology* 43, 942–948.
- Jensen ES (1996) Compared cycling in a soil-plant system of pea and barley residue nitrogen. *Plant and Soil* 182, 13–23.
- Kaye J, Finney D, White C, Bradley B, Schipanski M, Alonso-Ayuso M, Hunter M, Burgess M and Mejia C (2019) Managing nitrogen through cover crop species selection in the U.S. Mid-Atlantic. PLoS ONE 14, 1–23.
- Landis D, Menalled FD, Costamagna AC and Wilkinson TK (2005) Manipulating plant resources to enhance beneficial arthropods in agricultural landscapes. Weed Science 53, 902–908.
- **Leithold G, Hülsbergen KJ and Brock C** (2014) Organic matter returns to soils must be higher under organic compared to conventional farming. *Journal of Plant Nutrition and Soil Science* **178**, 4–12.
- Marcinkonis S, Pranaitis K and Lisova R (2007) Studies on use of various buckwheat biomasses. Lithuanian Institute of Agriculture, Voke Branch.
- Marshall S (2006) Insects: Their Natural History and Diversity: With a Photographic Guide to Insects of Eastern North America. Richmond Hills, Canada: Firefly Books.
- Martínez-Uña A, Martín JM, Fernández-Quintanilla C and Dorado J (2014)
  Provisioning floral resources to attract aphidophagous hoverflies (Diptera: Syrphidae) useful for pest management in Central Spain. *Journal of Economic Entomology* 106, 2327–2335.
- Morris JB, Chase C, Treadwell D, Koenig R, Cho A, Morales-Payan JP, Murphy T and Antonious GF (2015) Effect of sunn hemp (*Crotalaria juncea* L.) cutting date and planting density on weed suppression in Georgia, USA. *Journal of Environmental Sciences Health—Part B Pesticides, Food Contaminants, and Agricultural Wastes* 50, 614–621.
- NOAA (2013) Regional climate trends and scenarios for the U.S. national climate assessment: part 3. Climate of the Midwest U.S. NOAA Technical

- Report NESDIS 142-3. Available at https://scenarios.globalchange.gov/sites/default/files/NOAA\_NESDIS\_Tech\_Report\_142-3-Climate\_of\_the\_Midwest\_U.S\_0.pdf.
- NOP (2021) Organic production and handling requirements. Electronic Code of Federal Regulations. https://www.ecfr.gov/cgi-bin/text-idx?SID= 0b2eb9240015bee0bea089415abae567&mc=true&node=sp7.3.205.c&rgn= div6#se7.3.205\_1205.
- O'Connor R (2013) Reversing insect pollinator decline. *PostNote* 442, 1–5. Okansen J, Guillaume Blanchet F, Friendly M, Kindt R, Legendre P, McGlinn D, Minchin PR, O'Hara RB and Simpson GL (2019) Package 'vegan'. Available at https://cran.r-project.org/web/packages/vegan.
- O'Rourke ME, Liebman M and Rice ME (2008) Ground beetle (Coleoptera: Carabidae) assemblages in conventional and diversified crop rotation systems. *Environmental Entomology* 37, 121–130.
- Parr M, Grossman JM, Reberg-Horton SC, Brinton C and Crozier C (2011) Nitrogen delivery from legume cover crops in no-till organic corn production. Agronomy Journal 103, 1578–1590.
- Pender J and Mertz O (2006) Soil fertility depletion in sub-Saharan Africa: What is the role of organic agriculture?. In Halberg N, Alrøe HF, Knudsen MT and Kristensen ES (eds), Global Development of Organic Agriculture: Challenges and Prospects. CAB International, pp. 215–240. doi: 10.1079/9781845930783.0215.
- Piotrowska-Długosz A and Wilczewski E (2020) Influence of field pea (Pisum sativum L.) as catch crop cultivated for green manure on soil phosphorus and P-cycling enzyme activity. Archives of Agronomy and Soil Science 66, 1570–1582.
- Pulleman MM, Bouma J, van Essen EA and Meijles EW (2000) Soil organic matter content as a function of different land use history. Soil Science Society of America Journal 64, 689–693.
- R Core Team (2019) R: A Language and Environment for Statistical Computing. Vienna, Austria: R Foundation for Statistical Computing. Available at https://www.R-project.org/.
- Redlich S, Martin EA and Steffan-Dewenter I (2018) Landscape-level crop diversity benefits biological pest control. *Journal of Applied Ecology* 55, 2419–2428.
- Reicosky DC and Forcella F (1998) Cover crop and soil quality interactions in agroecosystems. *Journal of Soil and Water Conservation* **53**, 224–229.
- Robertson GP, Sollins P, Ellis BG and Lajtha K (1999) Exchangeable ions, pH, and cation exchange capacity. In Robertson GP, Coleman DC, Bledsoe CS and Sollins P (eds), Standard Soil Methods for Long-Term Ecological Research. Oxford: Oxford University Press, pp. 106–114.
- R Studio Team (2018) R Studio: Integrated Development Environment for R. Boston, MA: R Studio, Inc. Available at http://www.rstudio.com/.

- Sarrantonio M (2007) Building soil fertility and tilth with cover crops. In Clark A (ed.), *Managing Cover Crops Profitably*, 3rd Edn. Beltsville, MD: Sustainable Agricultural Network, pp. 16–24.
- Schiltz S, Munier-jolain N, Jeudy C, Burstin J and Salon C (2005) Dynamics of exogenous nitrogen partitioning and nitrogen remobilization from vegetative organs in pea revealed by 15N in vivo labeling throughout seed filling. *Plant Physiology* 137, 1463–1473.
- Schipanski ME, Barbercheck M, Douglas MR, Finney DM, Haider K, Kaye JP, Kemanian AR, Mortensen DA, Ryan MR, Tooker J and White C (2014) A framework for evaluating ecosystem services provided by cover crops in agroecosystems. *Agricultural Systems* 125, 12–22.
- Seitz S, Goebes P, Puerta VL, Pereira EIP, Wittwer R, Six J, van der Heijden MGA and Scholten T (2019) Conservation tillage and organic farming reduce soil erosion. Agronomy for Sustainable Development 39, 1–10.
- Sharma P, Singh A, Kahlon CS, Brar AS, Grover KK, Dia M and Steiner R (2018) The role of cover crops towards sustainable soil health and agriculture—A review paper. *American Journal of Plant Sciences* **09**, 1935–1951.
- Stepanovic SV, Burr C, Peterson JA, Rudnick D, Creech CF and Werle R (2018) Field pea response to seeding rate, depth, and inoculant in west-central Nebraska. *Agronomy Journal* 110, 1412–1419.
- Sumiahadi A, Acar R, Koç N and Özel A (2019) The potential use of cover crops for controlling soil erosion caused by water [Paper presentation]. 2019 International Soil Congress—Successful Transformation toward Land Degradation Neutrality: Future Perspective, Ankara, Turkey. Available at <a href="https://www.researchgate.net/publication/334119687\_The\_Potential\_Use\_of\_Cover\_Crops\_for\_Controlling\_Soil\_Erosion\_Caused\_by\_Water.">https://www.researchgate.net/publication/334119687\_The\_Potential\_Use\_of\_Cover\_Crops\_for\_Controlling\_Soil\_Erosion\_Caused\_by\_Water.</a>
- **Tilman D, Cassman KG, Matson PA, Naylor R and Polasky S** (2002) Agricultural sustainability and intensive production practices. *Nature Publishing Group* **418**, 671–677.
- Triplehorn CA, Johnson NF and Borror DJ (2005) Borror and DeLong's Introduction to the Study of Insects. Belmont, CA: Thompson Brooks/Cole.
- Wäckers FL, Romeis J and van Rijn P (2007) Nectar and pollen feeding by insect herbivores and implications for multitrophic interactions. *Annual Review of Entomology* **52**, 301–323.
- Wauters VM, Grossman J, Pfeiffer A and Cala R (2021) Ecosystem services and cash crop tradeoffs of summer cover crops in Northern region organic vegetable rotations. *Frontiers* 5, 1–11.
- Wilson H, Wong JS, Thorp RW, Miles AF, Daane KM and Altieri MA (2018) Pollinator ecology and management summer flowering cover crops support wild bees in vineyards. Environmental Entomology 47, 63–69.
- Zhang X, Davidson EA, Mauzerall DL, Searchinger TD, Dumas P and Shen Y (2015) Managing nitrogen for sustainable development. *Nature* **528**, 51–59.