Resilient urbanization for water limited environments

G. Darrel Jenerette
Center for Conservation Biology
Department of Botany and Plant Sciences
University of California Riverside
Riverside, CA USA

10.1017/dry.2024.7

Impact Statement

Dryland cities are home to more than two billion residents and are rapidly increasing throughout the world. This paper provides an overview of key challenges, opportunities, and uncertainties that affect the resilience of urban systems in drylands. New research directed to dryland cities is needed to help achieve sustainable development goals and contribute to more comprehensive theories of social-environmental systems spanning neighborhood to global scales.

Abstract

Dryland cities are important locations for human-environmental interactions and differ in many key characteristics from cities in wetter environments. The resilience of dryland cities depends on interactions across the entire urban continuum, from urban cores and suburban areas to teleconnected zones and wildland-urban interfaces. Defined by chronic water deficit, these cities face challenges that include securing essential resources, reducing vulnerability to hazards, and conserving threatened species. Resilience solutions must enhance the well-being of residents and institutions while fostering adaptive capacity throughout the urban continuum. Key axes of solutions include hydrologic integration, including stormwater capture and reuse, nature-based solutions, including expanding urban tree cover for cooling and health benefits, and landscape sustainability, including the incorporation of spatial heterogeneity into planning and development. Addressing the large uncertainties in ensuring more resilient cities requires convergence research, the integration of theoretically driven science that brings researchers and stakeholders together to identify problems, solutions, and opportunities for action. While convergence approaches look to address pressing scientific uncertainties they also are inherently place-based and address compelling case-studies to understand system dynamics and improve decision-making and land management. New research is needed to address the trade-offs resulting from decision-making and urban management activities, to meet the needs of diverse stakeholders, and to ensure that policies do not marginalize underserved communities. By leveraging innovative technologies, sustainable practices, and community involvement, dryland cities can overcome the challenges posed by chronic water limitations and thrive in their diverse environments.

Introduction

Cities are globally distributed key locations of human impacts to the biosphere and focal locations where people experience the environment (Grimm et al. 2008; Seto et al. 2012a). As complex socio-environmental systems, cities function through combinations of interactions that reflect self-organization, decision making, and external constraints (Bettencourt et al. 2007; Batty 2008). Cities are also highly heterogeneous landscapes with geographic variability spanning from the urban core featuring dense urban development and populations to associated teleconnected zones that provide resources for the urban core (Luck and Wu 2002; Zhou et al. 2017; Seto et al. 2012b). Given the increasing importance of urbanization to society and biosphere, extensive ongoing basic research and increasingly use-inspired research developed in collaboration with stakeholders are expanding to meet the theoretical and practical uncertainties of urban trajectories (Bai et al. 2017). Drylands are especially important loci of urbanization that differ in important characteristics from cities in wetter environments. With good reason, drylands served as the cradle of urban development in Mesopotamia, Indus Valley, China, and Mesoamérica as social organization was required for constructing elaborate irrigation systems (Wittfogel 1957; Scarborough 2003). Currently, 2.1 billion people or 56% of all urban residents live in a city where precipitation is substantially less than potential evaporation (Figure 1) and some of the most rapid rates of urbanization are occurring in dryland environments of Africa, Latin America, and Asia (Seto et al. 2012a; Liu et al. 2020). With global expansion of drylands and increasing frequency of dry conditions in historically wet regions (Huang et al. 2016), dryland cities also serve as a useful model for urban development and resilience globally. Exemplifying this challenge, by 2050 between 1.7-2.4 billion urban residents are projected to experience water scarcity (He et al. 2021). To achieve sustainable development goals and enhance opportunities for urban residents an expanded perspective of urbanized drylands is needed (FAO 2022).

Achieving sustainable development goals in the context of urbanization includes the maintenance of resources and environmental conditions for future generations while also enhancing an adaptive capacity to allow cities to continue responding to changing conditions (Wu 2014). Essential to both is increasing the resilience capacity of urbanization to ensure continued supply of resources, maintain natural biodiversity, and ensure equitable societal improvements (Chelleri et al. 2015; Meerow et al. 2016). Urban planning, from a resilience perspective, in part focuses on enhancing the capacity to recover and reorganize after disturbances such as fires and flooding from a biophysical driver or economic shocks and riots from societal drivers (Ahern 2011). The resilience perspective reflects both system- and landscape-attributes central to urban dynamics (Wu 2014; Groffman et al. 2017). Increasingly researchers and practitioners recognize that urban resilience capacity is dependent on interactions occurring throughout the urban continuum from urban core and suburban to more zones distal teleconnected and wildland urban interfaces (WUIs) (Seitzinger et al. 2012). Throughout these zones of urbanization innovation that generates urban adaptive capacity can be promoted (Ernstson et al. 2010). Fostering resilience in dryland cities is an essential goal for research, planning, and policy development to ensure progress in improving in achieving sustainable development goals.

To date, the majority of research into dryland cities and pathways to enhance their resilience has been primarily through case studies although some efforts have also been directed to cross-city comparisons. Phoenix, AZ USA in particular has served as an extensive test-bed to understand urban dryland dynamics and develop frameworks for enhancing their resilience. Beginning in 1998, the Central Arizona – Phoenix Long Term Ecological Research project has catalyzed researchers and practitioners throughout the metropolitan region. Key findings from this project have demonstrated the role of vegetation and associated irrigation on reducing neighborhood heat risks and that the distribution of vegetation density and cooler microclimates are associated with neighborhood wealth and racial makeup (Jenerette et al. 2008). This work has also pioneered research into urban resident desires leading to scenario framing as a pathway towards convergence research (Larson et al. 2019; Iwaniac et al. 2020; Brown et al. 2020). Other case study cities have also provided valuable insights into dryland urbanization, including ongoing research in Los Angeles, USA (McPherson et al. 2011; Kucera and Jenerette 2023), Barcelona, Spain (Baró et al. 2016; Anguelovski et al. 2018), Mexico City, Mexico (Velasco et al. 2005; Bravo-Bello et al. 2020), and Melbourne, Australia (Ossola et al. 2015; Threlfall et al. 2022). In addition to detailed case-studies, work has also been directed to cross site comparisons of dryland cities. Network approaches have been used in the southwestern United States to document the importance of climate and irrigation gradients on inter-urban variation in heat and vegetation (Jenerette et al. 2013; Shiflett et al. 2017; Ibsen et al. 2023). Global syntheses have also been conducted for dryland cities that evaluate sources of variation in urban heat risks (Wheeler et al. 2019; Dialesandro et al. 2019). Other comparative studies have included dryland cities in a broader selection of cities that have shown important differences between dryland cities compared to more mesic cities in context of vegetation, climate, and soils (Jenerette et al. 2016; Herrmann et al. 2020; Ibsen et al. 2021). With this baseline of cross city comparisons and on-going case studies there is a need for an overarching perspective of dryland urbanization. In looking toward this need, the objective here is to develop a framework for characterizing key sources of variability in dryland urbanization structure, function, and capacity for resilience. The resulting framework will be used to better identify challenges, solutions, and research needs to improve a theory of dryland cities and enhance the resilience of these systems.

Defining dryland urbanization and sources of variability

What is a water limited city or dryland urbanization? A defining feature of urbanization in drylands is a chronic terrestrial deficit between precipitation the capacity for evaporation (defined as an aridity index of 1 - precipitation / potential evapotranspiration less than 0.65), a commonly used definition of dryland (He et al. 2021). In this sense, a dryland city is one where the land surface could have higher rates of evaporation if additional water were available in an unmanaged condition. Characteristics reflecting an increasing magnitude of water deficit include infrequent precipitation events, low atmospheric humidity, high temperatures, dry soils, and reduced vegetation. In part because of these characteristics, dryland cities are exposed to elevated risks from water insecurity, wildfires, heatwaves, and flooding during episodic precipitation events (Lian et al. 2021; Yin et al. 2023). These features affect urban development and shape opportunities to enhance urban resilience in drylands. Nevertheless, this definition does not imply that dryland cities themselves lack water: in contrast many cities in dryland environments have extensive water resources. Water for urban uses can be acquired from surface and groundwater resources naturally located within or adjacent to the city (McDonald et al. 2014). Water is also captured from conveyances that may store and transport water great distances. Examples of such conveyance systems include the historic aquaducts of Rome that transported water more than 90 km and the modern Great Man-Made River in Libya that transports water up to 1,600 km in supplying multiple dryland cities. Thus, while the definition of terrestrial water limitation provides a key lens for considering dryland urban dynamics, the amount of water available to a city may be extensive and support additional vegetation, high rates of evaporation, and perennial downstream water flows.

In addition to the challenges of defining urban water limitation, identifying the scope of urbanization itself is also contested. Urbanization has been challenging to define as it is a combination of a system of interconnected social and environmental processes encompassing a large number of quasi-autonomous agents participating in webs of organizations with key processes occurring throughout a landscape (Wu 2014; Groffman et al. 2017). People and their interactions are the hallmark of urbanization where interactions span informal to formal as well as private to governmental institutions and interactions are characterized by varying degrees of cooperation and conflict. In this sense, urbanization encompasses a decentralized system of people and the environment. At the same time, urbanization is a geographic concept associated with extensive landscape heterogeneity found at multiple scales (Luck and Wu 2001; Seto et al. 2012b). Delineations of urban extent are connected to a variety of geographic indicators associated with population and social concentration, land cover distributions, governmental regulations (Raciti et al. 2012; Meerow et al. 2016). By using a simplified landscape model of urban development, a city can be separated into distinct zones of urbanization from the urban core, suburban, exurban, to teleconnected lands and wildland urban interfaces (Figure 1). Both among and within contrasting zones extensive variability in both landscape structure and system functioning are found. Within urban core areas, development is most intense with extensive distributions of impervious surfaces and limited amounts of cultivated vegetation, although many urban core areas include parks and other greenspaces. Within suburban areas, more open space is available, building materials diversify, and alternative opportunities for development are found. In exurban areas, the densities of built structures are low and extensive open space can be found. Finally, in teleconnected zones, the direct physical footprint of urban development is lacking but the effects of urbanization in these areas can lead to highly modified ecosystems for the production of goods, such as agriculture, or locations for waste (Seto et al. 2012b). Interfaces between urban and natural processes, or wildland urban interfaces (WUIs), are prominent interactions that further shape urban dynamics and can be found across all urbanization zones (Jenerette et al. 2022; Schug et al. 2023). Concomitant with this spatial variability, the different zones are integrated into the dynamics of the whole city. For example, while urban core and suburban zones frequently rely on imports of water resources, WUI, exurban and teleconnected zones may be sources of water that are brought to a city directly or transported virtually through other goods such as food (Jenerette and Larsen 2006; Paterson et al. 2015). Across all zones of urbanization both landscape and system characteristics influence urban structure and functioning. Together, a general definition of dryland urbanization is locations of urbanization occurring where the terrestrial landscape is water limited.

In addition to large heterogeneities within an urban system, cities differ dramatically throughout drylands. While a water balance definition of drylands suggests a threshold between dryland and more mesic cities it also reflects a continuum of cities from subhumid to hyper-arid conditions. These inter-urban differences further reflect differences in physiography, societies, and legacies of historical conditions. Differences in physiography can have a large effect on urban dynamics and sustainability challenges especially related to availability of water, heat, and open spaces. Climate differences among cities, including local water balance, timing of precipitation, or temperatures, have a large impact on current and future vulnerabilities (Imhoff et al. 2010; Anguelovski et al. 2016). Topographic variation, including presence of mountains or riverine systems can provide opportunities for enhanced water availability (Padowski and Gorelick 2014). Coastal conditions, either adjacent to an ocean or lake system, can also influence climatic conditions among cities and introduce large climate gradients within a single urban system (Tayyebi and Jenerette 2016). In contrast to physical drivers of inter-urban variation, social drivers also lead to large differences in urban structure and functioning, but the effects of these differences are not well understood. Cultural differences among dryland urban residents globally lead to contrasting urban design aesthetics and approaches for addressing sustainability challenges (Kihato et al. 2010). Social resources for enacting changes also vary among cities with wealthy cities having much greater capacity to affect change throughout contrasting urban zones and respond to extreme events as well as worsening chronic conditions (Zenghelis 2017). Finally, historical legacies differently influence dryland cities both in terms of structure and function (Schell et al. 2020; Roman et al. 2018). These sources of differences among cities individually and interactively can affect how different cities achieve more sustainable conditions that limit vulnerabilities of urban residents and institutions.

Challenges for Advancing Dryland Urban Resilience

Increasing opportunities to maintain and improve resident well-being and urban functioning is a crosscutting theme for achieving multiple sustainable development goals. While all urban systems face a wide range of sustainability challenges, cities in water limited environments have constraints that differ from more mesic environments and are exacerbated by increasing magnitudes of water deficit. These challenges include ensuring access to needed resources, reducing vulnerabilities to hazards, and ensuring conservation of threatened and endangered species. All of these challenges are influenced by multiple effects of urbanization on soils, vegetation, climate, and hydrology (Maldonado et al. 2023). Central to achieving sustainable urban futures is enhancing the resilience capacity of urban systems to cope with continually changing social and environmental conditions (Ernstson et al. 2010; Ahern 2011). Equitable distributions of urban conditions and the processes generating these conditions is a key challenge for all cities and reflects current decisions and legacies of systemic racism (Pulido 2000; Schell et al. 2020).

Elevated equity challenges in dryland cities may reflect the heightened dependence on key resources for well-being and the intensity of potential hazards (Jenerette et a. 2011). The process of crafting urban policy and management actions needs to leverage the diversity of conditions throughout a city and the corresponding sensitivities to environmental conditions.

Ensuring water security is a major constraint for dryland cities and is a focus of extensive urban planning (Hoekstra et al. 2018; McDonald et al. 2014; He et al. 2021). Water security encompasses multiple dimensions spanning uses and sources that vary spatially and temporally. Indoor water use is needed for residential, commercial, and industrial uses. Water is also used outdoors to support vegetation in both private and public green spaces as well as extensive water features such as artificial lakes or other recreational locations (Steele et al. 2014). In addition to water quantity, ensuring water quality criteria are satisfied is essential for ensuring its safe use. However, urbanization itself is a cause of degraded water quality associated in part with increasing contamination by novel synthetic compounds (McGrane 2016). To address water quality concerns, urban water undergoes a range of treatments before use, which varies with potable water requiring much higher standards than outdoor uses. Expanding equity in access to water in both contexts of quantity and quality is essential (Hoekstra et al. 2018; He et al. 2021). A widespread component of water inequity is associated with outdoor water uses for growing vegetation (Jenerette et al. 2011; Fang et al. 2023). In visioning for dryland urban resilience, the needs and strategies for ensuring adequate and equitable water resources varies within and throughout the different zones of urbanization. Throughout these zones the relative proportion of indoor to outdoor water will vary, for example urban core areas needing more water for residences and industry while in suburban areas the proportion of outdoor water may be larger. Similarly exurban, teleconnected, and WUI components may be sources of water resources and more suburban and core components much more consumers of water.

Sustainable access to additional resources beyond water, including energy and food as well as recreational opportunities and aesthetic benefits, are also needed for dryland cities. Frequently, these resources in dryland environments have substantial embodied or virtual water requirements (Garcia et al. 2020). Energy availability is a key resource with multiple ties to water availability: hydropower from both WUIs and teleconnected regions is a critical energy source while at the same time energy use for delivering water can be extensive either for water can require extensive energy for pumping and conveyance of freshwater or desalination of high saline water (Gober 2010; Wakeel et al. 2016). Potentially offsetting the energy demands of dryland cities, the generally clear sky conditions associated with drylands create favorable conditions for expanding solar energy production. Water in dryland environments is also used for food production, which occurs throughout urban zones from the urban core via urban gardening to teleconnected agricultural centers. Along with these “first order” resource needs urban residents also have needs for recreational and well-being activities that include access to greenspaces. These natural resources generally also require water resources. Ensuring adequate resources for urban systems especially in water limited environments where additional resources are also strongly connected to water availability is a key challenge in achieving resilience goals.

In addition to ensuring adequate resources, urban sustainability is also challenged by risks from hazards that are exacerbated by water limited conditions. For many dryland cities, fire in the WUI is an overarching concern that causes extensive damage directly through burning (Radeloff et al. 2018; Jones et al. 2022). Fires also contribute to degraded air quality, which can affect entire metropolitan regions. In addition to fires, high heat, especially occurring during heat waves are a major hazard to human health that is increasing in frequency, duration, and intensity and may be exacerbated by conditions associated with water limitation and urban development (Kalnay and Cai 2003; Patz et al. 2005; Maldonado et al. 2023). High heat is already the greatest cause of weather-related mortality and causes substantially more heat related illnesses (Vaidyanathan et al. 2020). High heat also compounds the effects of low air quality, both by increasing pollutants, for example with ozone formation and susceptibility (Sha et al. 2021; Areal et al. 2022). Flooding risks can also be high in water limited systems where soils may have limited water holding capacity and episodic nature of precipitation can cause flash flooding events (Kundzewicz et al. 2014; Yin et al. 2023). While not all hazards are exacerbated by increasing water limitation, many hazards are, and these pose critical challenges to the resilience of dryland cities.

Another challenge for many dryland cities is ensuring the sustainability of threatened and endangered species as well as rare habitats (Ren et al. 2022). Many dryland cities throughout the world are situated in areas with high conservation concern. Terrestrial species are directly affected by the loss and alteration of habitats associated with urban expansion with extensive risks occurring in WUIs as well as in exurban and teleconnected urban zones (McDonald et al. 2020). Aquatic species can be affected by development patterns but are also affected by hydrological alterations to water flow regimes and water quality (Cassady et al. 2023). Both terrestrial and aquatic species can be affected by releases of invasive species that often have origins of invasion within urban settings. Ensuring the persistence of natural biodiversity in the face of urban development is an ongoing sustainability challenge.

Solution axes for sustainable dryland urbanization

In addressing the sustainability challenges of cities in drylands, solutions are needed that enhance the safety and well-being of urban residents and institutions while fostering systems with robust adaptive capacity throughout the urban continuum. While individual challenges pose unique sources of uncertainty and require tailored planning to achieve specific goals, axes of solutions provide a more general approach for envisioning future dryland urban sustainability. The solution axes perspective highlights the importance of interactions among challenges and the reality that many challenges reflect a nexus of interconnected components (Liu et al. 2018). Across all solutions the combination of systems and landscape thinking is needed to evaluate the feedback between social and environmental dynamics that define urbanization.

One overarching solution axis is directed to ensuring hydrologic integration (Hoekstra et al. 2018). This has direct implications for ensuring water security, as well as the close linkages between water and energy, food production, and hazards including flooding, wildfire, and high heat (Wakeel et al. 2016). Water security is affected by both availability and usage; increasing water security can be achieved by appropriately modifying both. Diversifying sources of water, augmenting local water resources, and recycling water is a key direction for enhancing water security in dryland cities. Ensuring opportunities to capture and store local water can serve to both increase water resources while also reducing flood risks (Porse et al. 2018). Increasingly, especially for coastal cities but also inland cities with saline aquifers desalinization technologies are new sources of water, although this has large energy costs and potential environmental degradation to marine habitats through brine releases (Jones et al. 2019). Across the urban continuum there are many conflicts in water distributions especially between the urban core and suburban zones that use water for residential and commercial uses with those in exurban and teleconnected zones – primarily associated with agricultural water uses (Flörke et al. 2018). Improved hydrologic frameworks can help avoid these conflicts. One useful direction includes accounting for the water use efficiency of specific uses for comparison and including the context of embodied water (Shashua-Bar et al. 2009; Flörke et al. 2018). Addressing hydrologic issues as an integrated axis of dryland urbanization provides a comprehensive approach for addressing direct impacts, through water security and flooding control, while also indirectly affecting other resources, hazards, and well-being.

A second solution axis is the expansion of nature-based opportunities, also known as green infrastructure, or ecosystem services (Hobbie and Grimm 2020; FAO 2022; Fang et al. 2023). Nature-based solutions, such as enhancing vegetation, provide bundles of simultaneous services through suites of co-benefits (Raudsepp-Hearne et al. 2010; Lamy et al 2016). For example, nature-based solutions are frequently targeted to reduce risks such as heat, where vegetation cools surfaces through shading as well as cooling the larger areas through increased rates of evaporation (Gober et al. 2010; Kabisch et al. 2016; Iungman et al. 2023). A unique benefit of using vegetation for urban cooling is that this approach can function as a negative climate feedback, providing more cooling in hotter and drier conditions (Jenerette et al. 2011; Ibsen et al. 2021). For these reasons, in some dryland areas cities with irrigation supported vegetation can be cooler than outlying natural landscapes (Campos et al. 2022). At the same time, vegetation provides co-benefits of enhanced water infiltration and is often a pathway for reducing flooding risks and enhancing water quality (Berland et al. 2017). Finally, expanding vegetation in greenspaces also leads to many benefits for well-being associated with both mental and physical health (Rojas-Rueda et al. 2019). Highlighting these bundles of services, urban agriculture within the urban core or suburban zones is increasingly viewed as nature-based approach to enhance food security in cities while also providing cooling and health benefits (Brown and Jameton 2000). However, nature-based solutions approaches may not achieve all desired outcomes and may exacerbate other challenges. For example, while vegetation is a desired tool for reducing exposure to atmospheric pollutants extensive data suggests that the effects are minimal at best and can potentially enhance exposure to pollutants (Kumar et al. 2019). In another example, while increasing trees to shade buildings is desired for reducing high heat impacts, this approach can also elevate fire risks (Syphard et al. 2014). In the context of water use, a nature-based solutions axis for dryland cities will require increasing urban vegetation and outdoor water uses, which creates challenges for reducing water uses to enhance water security. An integrative and cross axis solution directed toward maximizing the water use efficiency of nature-based solutions is a needed approach for ensuring resilience of dryland cities.

Such an integrative approach is the hallmark of a third solution axis, a landscape sustainability approach, which can further incorporate hydrologic and nature-based solution axes to enhancing urban system resilience in drylands (Ahern 2012; Wu 2014; Wu 2021). Landscape sustainability science has been defined as a focus “on understanding how ecological consequences of spatial heterogeneity (e.g., species, community, ecosystem functions) cascade to affect human well-being (e.g., basic material, freedom of choice, health, social relations, security, inequality) as well as impacts of social changes to natural systems” (Liao et al. 2020). This approach explicitly features a design component and strong connections between managers, stakeholders, and researchers. Essential to a landscape sustainability approach is the characterization of location specific constraints and drivers of urban resilience. Importantly, a recognition that the constraints and drivers will vary across scales from individual parcels to entire metropolitan regions. Urban morphology and physiognomy can in many cases be strongly linked to sustainability indicators (Lamy et al. 2016; Zhang et al. 2023) because in part the location of landscape elements throughout a city can affect their functioning (Alberti 2005). Solutions that address and leverage spatial heterogeneity throughout cities provides a comprehensive axis for landscape and urban system management. Exemplifying the effects of landscape interactions, individual tree species can vary in their functional responses to the environment depending on where they are located (Ibsen et al. 2023), which is consistent with increases in the effectiveness of vegetation cooling increases in more arid locations (Tayyebi and Jenerette 2016). Harnessing cross-scale interactions provides additional mechanisms for balancing trade-offs. In particular, the inclusion of teleconnected zones as part of urbanization recognizes the inherent connectivity of urban and hinterland systems. Similarly, the inclusion of WUIs as part of the urban system opens opportunities to address key hazards, ecosystem services, and species conservation that simultaneously affect urban sustainability (Jenerette et al. 2022). A landscape sustainability framework provides a geographical and system dynamic framework that can further be combined with other axes. For example, landscape assessments that evaluate urban greening contributions to heat reductions and water uses (Zhang et al. 2017; Liang et al. 2017). Beyond science, a landscape sustainability approach requires engagement with urban designers, architects, and planners who envision future urban configurations as well as decision makers who implement policies (Nassauer 2012; Wu 2014). An example of this approach is the creation of superblocks in Barcelona, Spain – where a set of three by three street blocks of dense urban core area is transformed through creation of pedestrian and biking walkways, increasing vegetation and reducing car traffic (Mueller et al. 2020). While integrative in principle, implementation of a landscape sustainability approach is challenging due to the diversity of agents and environmental conditions found throughout cities, nonetheless such a framework is useful direction for enhancing the sustainability of dryland urbanization.

Research Needs

Many theoretical and place-based uncertainties limit the capacity to enhance resilience of dryland cities. While many science needs are general to all cities regardless of local climate conditions, dryland cities have needs specifically related to water security and addressing the many challenges that increase with aridity (Table 1). Convergence research, which brings together interdisciplinary scientists and knowledge users (National Research Council 2014), provides a framework to both address fundamental uncertainties in dryland urban system dynamics and contribute to decision making and land management. Such use inspired research into how solutions to dryland urban challenges can be most effectively implemented needs both systems and landscape perspectives to realize their benefits and minimize negative trade-offs. In part this challenge reflects uncertainties in how decision making across the large number of stakeholders with contrasting values can benefit everyone (Iwaniec et al. 2020; Mansur et al. 2022). An overarching concern is: How to best manage needs and consequences simultaneously for services, hazards, and conservation throughout the urban continuum and reconcile the trade-offs across these dimensions (Jenerette et al. 2022)? Concomitantly, how can practices be enacted that enhance equity in both environmental conditions and processes for enacting policies that do not further marginalize underserved communities (Anguelovski et al. 2022)? Addressing these complementary uncertainties in dryland cities should simultaneously enhance the well-being of urban residents, ensure the viability of threatened and endangered species, while encouraging resilience throughout the urban continuum.

Within these overarching concerns, one key research need is to better characterize the societal and cultural causes of variability in urban landscape structures and system dynamics that relate to resilience of these systems at global scales. Such aspects include differences in landscape design, demographic segregation, environmental feedback, The vast majority of urban research has been conducted in the United States, European Union, and Australia with expanding science contributions also originating from China. Latin America, Africa, the Middle East region, and much of Asia are not well represented in current data or models (Ziter 2016; Nagendra et al. 2018). These areas are some of the most rapidly expanding urban regions in the world and face increasing aridity. Cultural, socioeconomic, environmental, and historical characteristics in cities from these regions certainly differ from conditions where most research has been conducted. Characterizing the differences and causes among cities will lead to a more comprehensive consideration of urbanization and its interactions with local, adjacent, and teleconnected environments for places undergoing the most rapid changes.

In another research priority, while much research has been directed to urban sustainability within the urban core or suburban zones, more research directed to increased integration of urban sustainability with adjacent, i.e. WUIs, and teleconnected zones is needed (Seto et al. 2012; Jenerette et al. 2022). These more distal components of the urban system have strong impacts to urban dynamics associated with hazards, provisioning of key ecosystem services, and conservation of threatened and endangered species. Challenges with incorporating these distal urban components into frameworks of sustainable planning arise from the contrasting management jurisdictions, environments, and social systems between more central and distal urban zones. For example, WUIs are made of a patchwork of regulation and decision making that is frequently disconnected from the purview of urban core and suburban regions (Jenerette et al. 2022). Teleconnected zones can be distributed globally, limiting opportunities for direct management especially through urban policy making. While core-distal issues affect urbanization in all climatological conditions, in drylands key issues come to the fore, especially in response to fire and water security. For example, uncertainties in dryland WUI fire dynamics include both biophysical drivers of fire dynamics in response to both changing climate and land use patterns as well as public health issues related to effects of smoke exposure (D'Evelyn et al. 2022). Similarly, research is needed into water conveyance systems from WUI and teleconnected zones to suburban and urban core zones. Ensuring sustained water availability has both hydrological components into terrestrial water balances but also decision making across potentially competing uses (Jenerette and Larsen 2006; Flörke et al. 2018). These examples of fire and water security highlight the uncertainties in linking urban core with more distal zones showcase the need for geographically integrated frameworks for enhancing the resilience of dryland urbanization.

A third research need is reducing uncertainties in designing more effective urban structure to reduce climate vulnerabilities. Cities are warming much faster than natural ecosystems and are also exposed to greater risks from climate change induced disturbances including heatwaves, flooding, and fires. Novel interventions to reduce vulnerability to climate extremes and research that provides guidance on their implementation are needed (Wu 2014). The efficacy and potential trade-offs of adaptation strategies are in many cases uncertain and are related to the role of biological and landscape variation on functioning of individual interventions (Georgescu et al. 2014). For example, plant trait variation across and within species as well as local environmental context may affect the cooling capacity and water use efficiency of urban trees (Rahman et al. 2020; Yu et al. 2020). Unanswered question include: what species are most effective and is biodiversity important now and in the future? Where and in what arrangement should trees be planted to enhance effectiveness? How can expanding tree cooling capacity in vulnerable neighborhoods to contribute to gentrification? More broadly, research that evaluates alternative strategies not only for effects on risks but continues to assess actual outcomes, such as human health and energy demands. Addressing climate vulnerability research needs in dryland cities is especially needed as risks associated with hotter and drier conditions are increasingly having debilitating consequences to urban residents and systems.

Conclusions

Drylands are the cradle of urbanization and a crucible for developing sustainable pathways for a rapidly growing and urbanizing planet. Characterizing the variability of dryland urbanization, recognizing key challenges, applying axes of solutions and developing science that meaningfully improves urban resilience while contributing to a more comprehensive theory of human environmental interactions is a pressing scientific challenge (Figure 3). Chronic water limitation both poses multiple sustainability challenges to cities and yet has also been a source of innovation and organization both historically and today. These differences reflect environmental and social determinants as well as reflecting unique characteristics derived from the self-organization of individual cities. Enhancing the resilience of dryland cities is an important goal that will require interdisciplinary scientists, designers and planners, as well as decision makers and managers coming together. Framing urbanization as a multiple-scale phenomenon spanning a continuum from urban core to teleconnected zones with extensive WUIs throughout these zones provides a comprehensive perspective for understanding urban development and functioning. Envisioning solutions as axes or a nexus of interlinked approaches is a fruitful strategy to accomplish specific goals while also enhancing the capacity of urban systems. Dryland urbanization requires a holistic and adaptive approach that balances human needs with environmental sustainability. By leveraging innovative technologies, sustainable practices, and community involvement, cities in water limited regions can thrive despite the climate challenges. With a history extending to the first cities and a current unprecedented increase of urban development, improved knowledge of dryland urbanization will contribute both to advancing theories of human-environmental interactions and improving the lives of billions of people.

Acknowledgements

I thank Fernando Maestre for his invitation to submit this article and feedback on ideas during its development. Robert Johson conducted the GIS analyses for Figure 1. The background for the graphical abstract was generated using DALL-E (OpenAI).

Financial Support

This work was in part supported by the National Science Foundation (EAR–1923150).

Conflict of Interest Statement

Conflicts of Interest: None

Figure Captions

Figure 1 – Global distribution of cities of more than 250,000 residents in drylands (aridity index <0.65). Aridity layer was obtained from Trabucco and Zomer 2009. City data were obtained from geonames by opendatasoft.

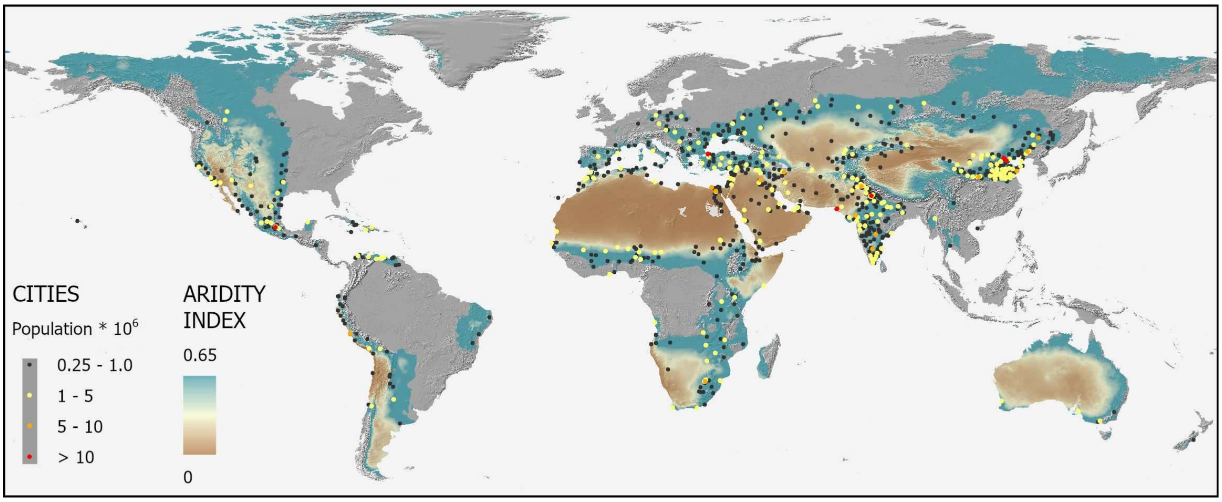
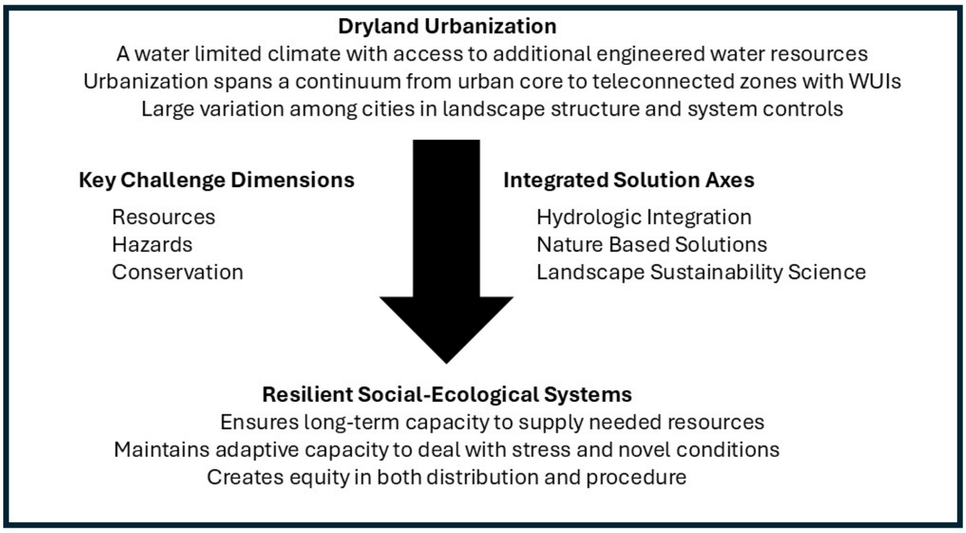


Figure 2 – Examples of contrasting zones of urbanization from urban core, suburban, exurban, and teleconnected lands (top to bottom) showing both urban and wildland urban interface (WUI) components (left and right). These examples are from the Los Angeles, CA USA metropolis. Imagery from Google Earth.

Figure 3 – Framework for organizing definitions, challenges, and solutions of dryland urbanization for achieving resilient social ecological systems. Dryland cities have multiple challenges that can reduce their capacity for resilience. Nevertheless, integrated solution axes instead of individual "fixes" provide pathways to address bundles of connected issues and address resulting tradeoffs.




Dryland
Urbanization
Challenges
Solutions
Use-Inspired
Research
Resilient
Social-Ecological
Systems

Tables

Table 1 – Key research needs for enhancing the resilience of urbanized drylands.

Overarching needs: How to best manage needs and consequences simultaneously for services, hazards, and conservation throughout the urban continuum and reconcile the trade-offs across these dimensions? Concomitantly, how can practices be enacted that enhance equity in both environmental conditions and processes for enacting policies that do not further marginalize underserved communities?

Major Research Challenge	Examples of Key Needs
1. Better characterize the causes of variability in dryland urbanization resilience	Expanding the geographic and cultural diversity of current research Identifying roles of cultural, socioeconomic, environmental, and historical characteristics
2. Increase integration of urbanization with adjacent wildland urban interfaces and teleconnected zones	Incorporating distal urban components into sustainable planning frameworks Characterizing WUI fire dynamics in response to changing climate and land use patterns and their public health consequences related to effects of smoke exposure Forecasting and improving water conveyance systems from distal components to suburban and urban core zones
3. Design more effective urban structure to reduce climate vulnerabilities	Identifying novel interventions to reduce vulnerability to climate extremes Evaluating how biological and landscape variation influences the efficacy and potential trade-offs of adaptation strategies Improving well-being without enhancing inequities or causing population displacement Evaluating alternative strategies for reducing vulnerabilities that directly assess desired actual outcomes

References

Ahern, J. 2011. From fail-safe to safe-to-fail: Sustainability and resilience in the new urban world. LANDSCAPE AND URBAN PLANNING 100:341–343.

Alberti, M. 2005. The effects of urban patterns on ecosystem function. INTERNATIONAL REGIONAL SCIENCE REVIEW 28:168–192.

Anguelovski, I., J. Connolly, H. Cole, M. Garcia-Lamarca, M. Triguero-Mas, F. Baró, N. Martin, D. Conesa, G. Shokry, C. del Pulgar, L. Ramos, A. Matheney, E. Gallez, E. Oscilowicz, J. Máñez, B. Sarzo, M. Beltrán, and J. Minaya. 2022. Green gentrification in European and North American cities. NATURE COMMUNICATIONS 13.

Anguelovski, I., L. Shi, E. Chu, D. Gallagher, K. Goh, Z. Lamb, K. Reeve, and H. Teicher. 2016. Equity Impacts of Urban Land Use Planning for Climate Adaptation: Critical Perspectives from the Global North and South. JOURNAL OF PLANNING EDUCATION AND RESEARCH 36:333–348.

Areal, A., Q. Zhao, C. Wigmann, A. Schneider, and T. Schikowski. 2022. The effect of air pollution when modified by temperature on respiratory health outcomes: A systematic review and meta-analysis. SCIENCE OF THE TOTAL ENVIRONMENT 811.

Bai, X., T. McPhearson, H. Cleugh, H. Nagendra, X. Tong, T. Zhu, and Y. Zhu. 2017. Linking Urbanization and the Environment: Conceptual and Empirical Advances. Pages 215–240 *in* A. Gadgil and T. Tomich, editors. ANNUAL REVIEW OF ENVIRONMENT AND RESOURCES, VOL 42.

Batty, M. 2008. The size, scale, and shape of cities. SCIENCE 319:769–771.

Berland, A., S. Shiflett, W. Shuster, A. Garmestani, H. Goddard, D. Herrmann, and M. Hopton. 2017. The role of trees in urban stormwater management. LANDSCAPE AND URBAN PLANNING 162:167–177.

Bettencourt, L., J. Lobo, D. Helbing, C. Kühnert, and G. West. 2007. Growth, innovation, scaling, and the pace of life in cities. PROCEEDINGS OF THE NATIONAL ACADEMY OF SCIENCES OF THE UNITED STATES OF AMERICA 104:7301–7306.

Brown, K., and A. Jameton. 2000. Public health implications of urban agriculture. JOURNAL OF PUBLIC HEALTH POLICY 21:20–39.

Brown, J.A., K.L. Larson, S.B. Lerman, D.L. Childers, R. Andrade, H.L. Bateman, S.J. Hall, P.S. Warren, and A.M. York. 2020. Influences of environmental and social factors on perceived bio-cultural services and disservices. FRONTIERS IN ECOLOGY AND EVOLUTION. DOI: 10.3389/fevo.2020.569730

Campos V. E., G. Gatica, N. Andino, V. N. F. Maldonado, A. Cardús. 2023. Land Surface Temperature in an Arid City: Assessing Spatio-temporal Changes. REMOTE SENSING IN EARTH SYSTEM SCIENCES 6:90–104.

Cassady, A., K. Anderson, K. Schwabe, and H. Regan. 2023. The Intersection of Wastewater Treatment Plants and Threatened and Endangered Species in California, USA Watersheds. WATER RESOURCES RESEARCH 59.

Chelleri, L., J. Waters, M. Olazabal, and G. Minucci. 2015. Resilience trade-offs: addressing multiple scales and temporal aspects of urban resilience. ENVIRONMENT AND URBANIZATION 27:181–198.

D'Evelyn, S., J. Jung, E. Alvarado, J. Baumgartner, P. Caligiuri, R. Hagmann, S. Henderson, P. Hessburg, S. Hopkins, E. Kasner, M. Krawchuk, J. Krenz, J. Lydersen, M. Marlier, Y. Masuda, K. Metlen, G. Mittelstaedt, S. Prichard, C. Schollaert, E. Smith, J. Stevens, C. Tessum, C. Reeb-Whitaker, J. Wilkins, N. Wolff, L. Wood, R. Haugo, and J. Spector. 2022. Wildfire, Smoke Exposure, Human Health, and Environmental Justice Need to be Integrated into Forest Restoration and Management. CURRENT ENVIRONMENTAL HEALTH REPORTS 9:366–385.

Dialesandro, JM, SM Wheeler, and Y Abunnasr. 2019. Urban Heat Island Behaviors in Dryland Regions. ENVIRONMENTAL RESEARCH COMMUNICATIONS 1:8 https://doi.org/10.1088/2515-7620/ab37d0.

Ernstson, H., S. van der Leeuw, C. Redman, D. Meffert, G. Davis, C. Alfsen, and T. Elmqvist. 2010. Urban Transitions: On Urban Resilience and Human-Dominated Ecosystems. AMBIO 39:531–545.

FAO. 2022. Urban forestry and urban greening in drylands - Improving resilience, health, and wellbeing of urban communities. A background document for the Green Urban Oases Programme. Rome. https://doi.org/10.4060/cc2065en

Fang, X., J. Li, and Q. Ma. 2023. Integrating green infrastructure, ecosystem services and nature-based solutions for urban sustainability: A comprehensive literature review. SUSTAINABLE CITIES AND SOCIETY 98.

Flörke, M., C. Schneider, and R. McDonald. 2018. Water competition between cities and agriculture driven by climate change and urban growth. NATURE SUSTAINABILITY 1:51–58.

Garcia, S., R. Rushforth, B. Ruddell, and A. Mejia. 2020. Full Domestic Supply Chains of Blue Virtual Water Flows Estimated for Major US Cities. WATER RESOURCES RESEARCH 56.

Georgescu, M., P. Morefield, B. Bierwagen, and C. Weaver. 2014. Urban adaptation can roll back warming of emerging megapolitan regions. PROCEEDINGS OF THE NATIONAL ACADEMY OF SCIENCES OF THE UNITED STATES OF AMERICA 111:2909–2914.

Gober, P. 2010. Desert urbanization and the challenges of water sustainability. CURRENT OPINION IN ENVIRONMENTAL SUSTAINABILITY 2:144–150.

Gober, P., A. Brazel, R. Quay, S. Myint, S. Grossman-Clarke, A. Miller, and S. Rossi. 2010. Using Watered Landscapes to Manipulate Urban Heat Island Effects: How Much Water Will It Take to Cool Phoenix? JOURNAL OF THE AMERICAN PLANNING ASSOCIATION 76:109–121.

Grimm, N., S. Faeth, N. Golubiewski, C. Redman, J. Wu, X. Bai, and J. Briggs. 2008. Global change and the ecology of cities. SCIENCE 319:756–760.

Groffman, P., M. Cadenasso, J. Cavender-Bares, D. Childers, N. Grimm, J. Grove, S. Hobbie, L. Hutyra, G. Jenerette, T. McPhearson, D. Pataki, S. Pickett, R. Pouyat, E. Rosi-Marshall, and B. Ruddell. 2017. Moving Towards a New Urban Systems Science. ECOSYSTEMS 20:38–43.

He, C., Z. Liu, J. Wu, X. Pan, Z. Fang, J. Li, and B. Bryan. 2021. Future global urban water scarcity and potential solutions. NATURE COMMUNICATIONS 12.

Herrmann, DL, LA Schifman, and WD Shuster. 2020. Urbanization Drives Convergence in Soil Profile Texture and Carbon Content. ENVIRONMENTAL RESEARCH LETTERS 15, no. 11. https://doi.org/10.1088/1748-9326/abbb00.

Hobbie, S., and N. Grimm. 2020. Nature-based approaches to managing climate change impacts in cities. PHILOSOPHICAL TRANSACTIONS OF THE ROYAL SOCIETY B-BIOLOGICAL SCIENCES 375.

Hoekstra, A., J. Buurman, and K. van Ginkel. 2018. Urban water security: A review. ENVIRONMENTAL RESEARCH LETTERS 13.

Huang, J., H. Yu, X. Guan, G. Wang, and R. Guo. 2016. Accelerated dryland expansion under climate change. NATURE CLIMATE CHANGE 6:166-171.

Ibsen, P., D. Borowy, T. Dell, H. Greydanus, N. Gupta, D. Hondula, T. Meixner, M. Santelmann, S. Shiflett, M. Sukop, C. Swan, M. Talal, M. Valencia, M. Wright, and G. Jenerette. 2021. Greater aridity increases the magnitude of urban nighttime vegetation-derived air cooling. ENVIRONMENTAL RESEARCH LETTERS 16.

Ibsen, PC, LS Santiago, SA Shiflett, M Chandler, and GD Jenerette. 2023. Irrigated Urban Trees Exhibit Greater Functional Trait Plasticity Compared to Natural Stands. BIOLOGY LETTERS 19:1 https://doi.org/10.1098/rsbl.2022.0448.

Imhoff, M., P. Zhang, R. Wolfe, and L. Bounoua. 2010. Remote sensing of the urban heat island effect across biomes in the continental USA. REMOTE SENSING OF ENVIRONMENT 114:504–513.

Iungman, T., M. Cirach, F. Marando, E. Barboza, S. Khomenko, P. Masselot, M. Quijal-Zamorano, N. Mueller, A. Gasparrini, J. Urquiza, M. Heris, M. Thondoo, and M. Nieuwenhuijsen. 2023. Cooling cities through urban green infrastructure: a health impact assessment of European cities. LANCET 401:577–589.

Iwaniec, D., E. Cook, M. Davidson, M. Berbés-Blázquez, M. Georgescu, E. Krayenhoff, A. Middel, D. Sampson, and N. Grimm. 2020. The co-production of sustainable future scenarios. LANDSCAPE AND URBAN PLANNING 197.

Jenerette, G., K. Anderson, M. Cadenasso, M. Fenn, J. Franklin, M. Goulden, L. Larios, S. Pincetl, H. Regan, S. Rey, L. Santiago, and A. Syphard. 2022. An expanded framework for wildland-urban interfaces and their management. FRONTIERS IN ECOLOGY AND THE ENVIRONMENT 20:516–523.

Jenerette, GD, G Miller, A Buyantuev, DE Pataki, T Gillespie, and S Pincetl. 2013. Urban Vegetation and Income Segregation in Drylands: A Synthesis of Seven Metropolitan Regions in the Southwestern United States." ENVIRONMENTAL RESEARCH LETTERS 8:4. https://doi.org/10.1088/1748-9326/8/4/044001.

Jenerette, G., S. Harlan, W. Stefanov, and C. Martin. 2011. Ecosystem services and urban heat riskscape moderation: water, green spaces, and social inequality in Phoenix, USA. ECOLOGICAL APPLICATIONS 21:2637–2651.

Jenerette, G., and L. Larsen. 2006. A global perspective on changing sustainable urban water supplies. GLOBAL AND PLANETARY CHANGE 50:202–211.

Jones, E., M. Qadir, M. van Vliet, V. Smakhtin, and S. Kang. 2019. The state of desalination and brine production: A global outlook. SCIENCE OF THE TOTAL ENVIRONMENT 657:1343–1356.

Jones, M., J. Abatzoglou, S. Veraverbeke, N. Andela, G. Lasslop, M. Forkel, A. Smith, C. Burton, R. Betts, G. van der Werf, S. Sitch, J. Canadell, C. Santín, C. Kolden, S. Doerr, and C. Le Quéré. 2022.

Global and Regional Trends and Drivers of Fire Under Climate Change. REVIEWS OF GEOPHYSICS 60.

Kabisch, N., N. Frantzeskaki, S. Pauleit, S. Naumann, M. Davis, M. Artmann, D. Haase, S. Knapp, H. Korn, J. Stadler, K. Zaunberger, and A. Bonn. 2016. Nature-based solutions to climate change mitigation and adaptation in urban areas: perspectives on indicators, knowledge gaps, barriers, and opportunities for action. ECOLOGY AND SOCIETY 21.

Kalnay, E., and M. Cai. 2003. Impact of urbanization and land-use change on climate. NATURE 423:528–531.

Kihato CW, M Massoumi, BA Ruble (eds). 2010. Urban Diversity: Space, Culture, and Inclusive Pluralism in Cities Worldwide. Woodrow Wilson Center Press with Johns Hopkins University Press

Kumar, P., A. Druckman, J. Gallagher, B. Gatersleben, S. Allison, T. Eisenman, U. Hoang, S. Hama, A. Tiwari, A. Sharma, K. Abhijith, D. Adlakha, A. McNabola, T. Astell-Burt, X. Feng, A. Skeldon, S. de Lusignan, and L. Morawska. 2019. The nexus between air pollution, green infrastructure and human health. ENVIRONMENT INTERNATIONAL 133.

Kundzewicz, Z., S. Kanae, S. Seneviratne, J. Handmer, N. Nicholls, P. Peduzzi, R. Mechler, L. Bouwer, N. Arnell, K. Mach, R. Muir-Wood, G. Brakenridge, W. Kron, G. Benito, Y. Honda, K. Takahashi, and B. Sherstyukov. 2014. Flood risk and climate change: global and regional perspectives. HYDROLOGICAL SCIENCES JOURNAL 59:1–28.

Larson, K.L., E.A. Corley, R. Andrade, S.J. Hall, A.M. York, S.A. Meerow, P.J. Coseo, D.L. Childers and D.M. Hondula. 2019. Subjective evaluations of ecosystem services and disservices: an approach to creating and analyzing robust survey scales. ECOLOGY AND SOCIETY. DOI: 10.5751/ES10888-240207.

Lamy, T., K. Liss, A. Gonzalez, and E. Bennett. 2016. Landscape structure affects the provision of multiple ecosystem services. ENVIRONMENTAL RESEARCH LETTERS 11.

Lian, X., S. Piao, A. Chen, C. Huntingford, B. Fu, L. Li, J. Huang, J. Sheffield, A. Berg, T. Keenan, T. McVicar, Y. Wada, X. Wang, T. Wang, Y. Yang, and M. Roderick. 2021. Multifaceted characteristics of dryland aridity changes in a warming world. NATURE REVIEWS EARTH & ENVIRONMENT 2:232–250.

Liang, L., R. Anderson, S. Shiflett, and G. Jenerette. 2017. Urban outdoor water use and response to drought assessed through mobile energy balance and vegetation greenness measurements. ENVIRONMENTAL RESEARCH LETTERS 12.

Liao, C, JX Qiu, B Chen, DL Chen, BJ Fu, M Georgescu, CY He, et al. 2020. Advancing Landscape Sustainability Science: Theoretical Foundation and Synergies with Innovations in Methodology, Design, and Application. LANDSCAPE ECOLOGY 35:1–9. https://doi.org/10.1007/s10980-020-00967-0.

Liu, J., V. Hull, H. Godfray, D. Tilman, P. Gleick, H. Hoff, C. Pahl-Wostl, Z. Xu, M. Chung, J. Sun, and S. Li. 2018. Nexus approaches to global sustainable development. NATURE SUSTAINABILITY 1:466–476.

Liu, X., Y. Huang, X. Xu, X. Li, X. Li, P. Ciais, P. Lin, et al. 2020. High-Spatiotemporal-Resolution Mapping of Global Urban Change from 1985 to 2015. Nature Sustainability 3:564–70. https://doi.org/10.1038/s41893-020-0521-x.

Luck, M., and J. Wu. 2002. A gradient analysis of urban landscape pattern: a case study from the Phoenix metropolitan region, Arizona, USA. LANDSCAPE ECOLOGY 17:327–339.

Maldonado, V. N. F., G. Gatica, A. Cardus, V. E. Campos. 2023. Evaluación de los cambios en el uso y cobertura del suelo en una ciudad en desarrollo basado en imágenes satelitales. REVISTA de GEOGRAPHÍA (SAN JUAN) 26:4-14.

Mansur, A., R. McDonald, B. Güneralp, H. Kim, J. de Oliveira, C. Callaghan, P. Hamel, J. Kuiper, M. Wolff, V. Liebelt, I. Martins, T. Elmqvist, and H. Pereira. 2022. Nature futures for the urban century: Integrating multiple values into urban management. ENVIRONMENTAL SCIENCE & POLICY 131:46–56.

McDonald, R., K. Weber, J. Padowski, M. Flörke, C. Schneider, P. Green, T. Gleeson, S. Eckman, B. Lehner, D. Balk, T. Boucher, G. Grill, and M. Montgomery. 2014. Water on an urban planet: Urbanization and the reach of urban water infrastructure. GLOBAL ENVIRONMENTAL CHANGE-HUMAN AND POLICY DIMENSIONS 27:96–105.

McGrane, S. 2016. Impacts of urbanisation on hydrological and water quality dynamics, and urban water management: a review. HYDROLOGICAL SCIENCES JOURNAL 61:2295–2311.

Meerow, S., J. Newell, and M. Stults. 2016. Defining urban resilience: A review. LANDSCAPE AND URBAN PLANNING 147:38–49.

Mueller, N., D. Rojas-Rueda, H. Khreis, M. Cirach, D. Andrés, J. Ballester, X. Bartoll, C. Daher, A. Deluca, C. Echave, C. Milà, S. Márquez, J. Palou, K. Pérez, C. Tonne, M. Stevenson, S. Rueda, and M. Nieuwenhuijsen. 2020. Changing the urban design of cities for health: The superblock model. ENVIRONMENT INTERNATIONAL 134:105132.

Nagendra, H., X. Bai, E. Brondizio, and S. Lwasa. 2018. The urban south and the predicament of global sustainability. NATURE SUSTAINABILITY 1:341–349.

Nassauer, J. 2012. Landscape as medium and method for synthesis in urban ecological design. LANDSCAPE AND URBAN PLANNING 106:221–229.

Padowski, J., and S. Gorelick. 2014. Global analysis of urban surface water supply vulnerability. ENVIRONMENTAL RESEARCH LETTERS 9.

Paterson, W., R. Rushforth, B. Ruddell, M. Konar, I. Ahams, J. Gironás, A. Mijic, and A. Mejia. 2015. Water Footprint of Cities: A Review and Suggestions for Future Research. SUSTAINABILITY 7:8461–8490.

Patz, J., D. Campbell-Lendrum, T. Holloway, and J. Foley. 2005. Impact of regional climate change on human health. NATURE 438:310–317.

Porse, E., K. Mika, E. Litvak, K. Manago, T. Hogue, M. Gold, D. Pataki, and S. Pincetl. 2018. The economic value of local water supplies in Los Angeles. NATURE SUSTAINABILITY 1:289–297.

Pulido, L. 2000. Rethinking environmental racism: White privilege and urban development in southern California. ANNALS OF THE ASSOCIATION OF AMERICAN GEOGRAPHERS 90:12–40.

Raciti, S., L. Hutyra, P. Rao, and A. Finzi. 2012. Inconsistent definitions of "urban" result in different conclusions about the size of urban carbon and nitrogen stocks. ECOLOGICAL APPLICATIONS 22:1015–1035.

Radeloff, V., D. Helmers, H. Kramer, M. Mockrin, P. Alexandre, A. Bar-Massada, V. Butsic, T. Hawbaker, S. Martinuzzi, A. Syphard, and S. Stewart. 2018. Rapid growth of the US wildland-urban interface raises wildfire risk. PROCEEDINGS OF THE NATIONAL ACADEMY OF SCIENCES OF THE UNITED STATES OF AMERICA 115:3314–3319.

Rahman, M., L. Stratopoulos, A. Moser-Reischl, T. Zölch, K. Häberle, T. Rötzer, H. Pretzsch, and S. Pauleit. 2020. Traits of trees for cooling urban heat islands: A meta-analysis. BUILDING AND ENVIRONMENT 170.

Raudsepp-Hearne, C., G. Peterson, and E. Bennett. 2010. Ecosystem service bundles for analyzing tradeoffs in diverse landscapes. PROCEEDINGS OF THE NATIONAL ACADEMY OF SCIENCES OF THE UNITED STATES OF AMERICA 107:5242–5247.

Ren, Q., C. He, Q. Huang, P. Shi, D. Zhang, and B. Güneralp. 2022. Impacts of urban expansion on natural habitats in global drylands. NATURE SUSTAINABILITY 5:869–878.

Rojas-Rueda, D., M. Nieuwenhuijsen, M. Gascon, D. Perez-Leon, and P. Mudu. 2019. Green spaces and mortality: a systematic review and meta-analysis of cohort studies. LANCET PLANETARY HEALTH 3:E469–E477.

Roman, L., H. Pearsall, T. Eisenman, T. Conway, R. Fahey, S. Landry, J. Vogt, N. van Doorn, J. Grove, D. Locke, A. Bardekjian, J. Battles, M. Cadenasso, C. van den Bosch, M. Avolio, A. Berland, G. Jenerette, S. Mincey, D. Pataki, and C. Staudhammer. 2018. Human and biophysical legacies shape contemporary urban forests: A literature synthesis. URBAN FORESTRY & URBAN GREENING 31:157–168.

Scarborough, Vernon L. 2003. "The Flow of Power: Ancient Water Systems and Landscapes." SAR Press.

Schell, C., K. Dyson, T. Fuentes, S. Des Roches, N. Harris, D. Miller, C. Woelfle-Erskine, and M. Lambert. 2020. The ecological and evolutionary consequences of systemic racism in urban environments. SCIENCE 369:1446-+.

Schug, F., A. Bar-Massada, A. Carlson, H. Cox, T. Hawbaker, D. Helmers, P. Hostert, D. Kaim, N. Kasraee, S. Martinuzzi, M. Mockrin, K. Pfoch, and V. Radeloff. 2023. The global wildland-urban interface. NATURE 621:94-+.

Seitzinger, S., U. Svedin, C. Crumley, W. Steffen, S. Abdullah, C. Alfsen, W. Broadgate, F. Biermann, N. Bondre, J. Dearing, L. Deutsch, S. Dhakal, T. Elmqvist, N. Farahbakhshazad, O. Gaffney, H. Haberl, S. Lavorel, C. Mbow, A. McMichael, J. deMoraís, P. Olsson, P. Pinho, K. Seto, P. Sinclair, M. Smith, and L. Sugar. 2012. Planetary Stewardship in an Urbanizing World: Beyond City Limits. AMBIO 41:787–794.

Seto, K., B. Güneralp, and L. Hutyra. 2012a. Global forecasts of urban expansion to 2030 and direct impacts on biodiversity and carbon pools. PROCEEDINGS OF THE NATIONAL ACADEMY OF SCIENCES OF THE UNITED STATES OF AMERICA 109:16083–16088.

Seto, K., A. Reenberg, C. Boone, M. Fragkias, D. Haase, T. Langanke, P. Marcotullio, D. Munroe, B. Olah, and D. Simon. 2012b. Urban land teleconnections and sustainability. PROCEEDINGS OF THE NATIONAL ACADEMY OF SCIENCES OF THE UNITED STATES OF AMERICA 109:7687–7692.

Sha, T., X. Ma, H. Zhang, N. Janechek, Y. Wang, Y. Wang, L. García, G. Jenerette, and J. Wang. 2021. Impacts of Soil NOx Emission on O3 Air Quality in Rural California. ENVIRONMENTAL SCIENCE & TECHNOLOGY 55:7113–7122.

Shashua-Bar, L., D. Pearlmutter, and E. Erell. 2009. The cooling efficiency of urban landscape strategies in a hot dry climate. LANDSCAPE AND URBAN PLANNING 92:179–186.

Shiflett, SA, LYL Liang, SM Crum, GL Feyisa, J Wang, and GD Jenerette. 2017. Variation in the Urban Vegetation, Surface Temperature, Air Temperature Nexus. SCIENCE OF THE TOTAL ENVIRONMENT 579: 495–505. https://doi.org/10.1016/j.scitotenv.2016.11.069.

Simkin, R., K. Seto, R. McDonald, and W. Jetz. 2022. Biodiversity impacts and conservation implications of urban land expansion projected to 2050. PROCEEDINGS OF THE NATIONAL ACADEMY OF SCIENCES OF THE UNITED STATES OF AMERICA 119.

Steele, M., and J. Heffernan. 2014. Morphological characteristics of urban water bodies: mechanisms of change and implications for ecosystem function. ECOLOGICAL APPLICATIONS 24:1070–1084.

Syphard, A., T. Brennan, and J. Keeley. 2014. The role of defensible space for residential structure protection during wildfires. INTERNATIONAL JOURNAL OF WILDLAND FIRE 23:1165–1175.

Tayyebi, A., and G. Jenerette. 2016. Increases in the climate change adaption effectiveness and availability of vegetation across a coastal to desert climate gradient inmetropolitan Los Angeles, CA, USA. SCIENCE OF THE TOTAL ENVIRONMENT 548:60–71.

Trabucco, A., and Zomer, R.J. 2009. Global Aridity Index (Global-Aridity) and Global Potential Evapo-Transpiration (Global-PET) Geospatial Database. CGIAR Consortium for Spatial Information. Published online, available from the CGIAR-CSI GeoPortal at: http://www.csi.cgiar.org.

Vaidyanathan, A., J. Malilay, P. Schramm, and S. Saha. 2020. Heat-Related Deaths - United States, 2004-2018. MMWR-MORBIDITY AND MORTALITY WEEKLY REPORT 69:729–734.

Wakeel, M., B. Chen, T. Hayat, A. Alsaedi, and B. Ahmad. 2016. Energy consumption for water use cycles in different countries: A review. APPLIED ENERGY 178:868–885.

Wheeler, SM, Y Abunnasr, J Dialesandro, E Assaf, S Agopian, and VC Gamberini. 2019. Mitigating Urban Heating in Dryland Cities: A Literature Review. JOURNAL OF PLANNING LITERATURE 34: 434–46. https://doi.org/10.1177/0885412219855779.

Wittfogel, K. A. 1957. Oriental Despotism: A Comparative Study of Total Power. Yale University Press.

Wu, J. 2014. Urban ecology and sustainability: The state-of-the-science and future directions. LANDSCAPE AND URBAN PLANNING 125:209–221.

Wu, J. 2021. Landscape sustainability science (II): core questions and key approaches. LANDSCAPE ECOLOGY 36:2453-2485.

Yin J., Y. Gao, R. Chen, D. Yu, R. Wilby, N. Wright, Y. Ge, J. Bricker, H. Gong, and M Guan. 2023. Flashfloods: why are more of them devastating the world's driest regions? NATURE 615:212-215.

Yu, Z., G. Yang, S. Zuo, G. Jorgensen, M. Koga, and H. Vejre. 2020. Critical review on the cooling effect of urban blue-green space: A threshold-size perspective. URBAN FORESTRY & URBAN GREENING 49.

Zenghelis, D. 2017. Equity and national mitigation. NATURE CLIMATE CHANGE 7:9–10.

Zhang, P., D. Ghosh, and S. Park. 2023. Spatial measures and methods in sustainable urban morphology: A systematic review. LANDSCAPE AND URBAN PLANNING 237.

Zhang, Y., A. Murray, and B. Turner. 2017. Optimizing green space locations to reduce daytime and nighttime urban heat island effects in Phoenix, Arizona. LANDSCAPE AND URBAN PLANNING 165:162–171.

Zhou, W., J. Wang, and M. Cadenasso. 2017. Effects of the spatial configuration of trees on urban heat mitigation: A comparative study. REMOTE SENSING OF ENVIRONMENT 195:1–12.

Ziter, C. 2016. The biodiversity-ecosystem service relationship in urban areas: a quantitative review. OIKOS 125:761–768.